

1     **Integrated hydrogeophysical modelling and data assimilation for geoelectrical**  
2                     **leak detection**

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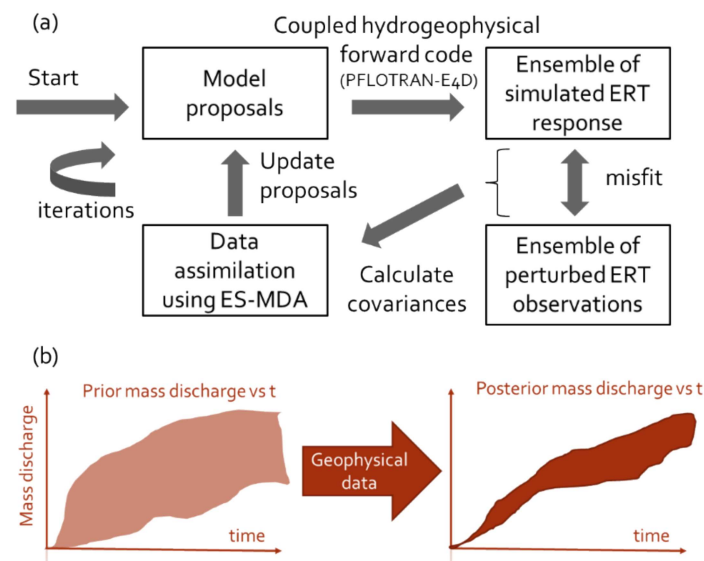
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## Abstract

3 Time-lapse electrical resistivity tomography (ERT) measurements provide indirect observations of hydrological  
 4 processes in the Earth's shallow subsurface at high spatial and temporal resolution. ERT has been used in the  
 5 past decades to detect leaks and monitor the evolution of associated contaminant plumes. Specifically, inverted  
 6 resistivity images allow visualization of the dynamic changes in the structure of the plume. However, existing  
 7 methods do not allow the direct estimation of leak parameters (e.g. leak rate, location, etc.) and their  
 8 uncertainties. We propose an ensemble-based data assimilation framework that evaluates proposed hydrological  
 9 models against observed time-lapse ERT measurements without directly inverting for the resistivities. Each  
 10 proposed hydrological model is run through the parallel coupled hydro-geophysical simulation code  
 11 PFLOTRAN-E4D to obtain simulated ERT measurements. The ensemble of model proposals is then updated  
 12 using an iterative ensemble smoother. We demonstrate the proposed framework on synthetic and field ERT data  
 13 from controlled tracer injection experiments. Our results show that the approach allows joint identification of  
 14 contaminant source location, initial release time, and solute loading from the cross-borehole time-lapse ERT  
 15 data, alongside with an assessment of uncertainties in these estimates. We demonstrate a reduction in site-wide  
 16 uncertainty by comparing the prior and posterior plume mass discharges at a selected image plane. This  
 17 framework is particularly attractive to sites that have previously undergone extensive geological investigation  
 18 (e.g., nuclear sites). It is well suited to complement ERT imaging and we discuss practical issues in its  
 19 application to field problems.

## Graphical abstract



20  
 21 **Estimation of leak parameters and their uncertainties using raw geophysical data and data assimilation.**

22

## 1) Introduction

23 Identification of solute loadings from an unknown source is a complex yet critical problem. For example,  
 24 understanding the whereabouts of the source(s) of contamination is often the first question that needs to be  
 25 addressed in a remediation project. This identification, however, is not straightforward and it is often  
 26 complicated by factors such as unknown forcing (e.g., boundary and flow conditions), aquifer and vadose zone  
 27 heterogeneity, and limited data (in terms of number, types, temporal and spatial coverage). Because of these

28 complications, attempts to assess source identification should also address the uncertainties in the estimates,  
29 and provide realistic and actionable uncertainty bounds.

30 Traditional point-based sampling methods suffer from limited coverage and resolution. As prompted,  
31 in part, by the wealth of studies in stochastic subsurface hydrology that argued for better field techniques,  
32 geophysical methods have emerged as valuable tools for investigating shallow subsurface processes over the  
33 past two decades (Binley et al., 2015). Geophysical methods can provide much larger spatial and temporal  
34 resolution. Once installed, autonomous long-term monitoring systems, such as ALERT (Kuras et al., 2009), can  
35 repeatedly collect geophysical data and transmit it back to the office using telemetry. Among them, electrical  
36 resistivity tomography (ERT) is particularly suitable for leak detection due to its sensitivity to fluid conductivity.  
37 Note that leak detection is not limited to the detection of the breakthrough of saline fluids (as proxies of  
38 contaminants), but it also includes monitoring the integrity of water-retaining structures (e.g. embankments or  
39 levees) (Abdulsamad et al., 2019) and landfills (Audebert et al., 2014; Chambers et al., 2006; Maurya et al.,  
40 2017). Landfill sites with a high content of inorganics tend to provide good signal (Maurya et al., 2017), which  
41 in some cases can be linked to other contaminants from the same source and following the same pathways.  
42 However, this is a critical assumption which needs to be tested at each site (Balbarini et al., 2018). Many previous  
43 ERT studies have focused on inferring plume characteristics by delineating the plume geometry (Aghasi et al.,  
44 2013), obtaining summary statistics of the plume structure (e.g. spatial and temporal movements) (Crestani et al.,  
45 2015; Pidlisecky et al., 2011; Singha and Gorelick, 2006), or developing methods for automatic tracking of  
46 plumes (Ward et al., 2016). There is also a substantial amount of work dedicated to delineating local hydraulic  
47 properties using ERT (e.g. Camporese et al., 2011). As an effort to better use geophysical data for  
48 hydrogeological studies, comparisons between coupled and uncoupled hydrogeophysical inversions of ERT-  
49 monitored tracer tests have been made (Camporese et al., 2015; Hinnell et al., 2010). Others have tried to address  
50 the uncertain link between hydrological systems and geophysical data using data-driven or machine learning  
51 approaches (Hermans et al., 2016, 2015; Oware et al., 2013). There is also increased use of geophysics to  
52 estimate remediation efficiency (LaBrecque et al., 1996). For example, Power et al. (2014) applied 4D active  
53 time-constrained inversion to time-lapse ERT data to estimate the volume of solute plume remediated in a  
54 laboratory experiment, while Slater and Binley (2006, 2003) used electrical imaging to monitor the integrity of  
55 permeable reactive barriers. Its applicability largely depends whether the plume (e.g. saline or inorganic) and  
56 the injected agents (e.g. zero valent iron or oxidants) or background gives distinct electrical signals. Plumes  
57 with non-charged compounds, such as chlorinated ethenes, tend not to give an ERT response, except at extreme  
58 concentrations. Likewise, in nuclear sites, the concentration of radionuclides itself tend not to generate a large  
59 enough signal but secondary species such as metallic ions may give a distinctive ERT response.

60 The various electrical methods applied to the mapping and monitoring at the U.S. Department of Energy  
61 Hanford nuclear site has greatly improved the readiness of these methods (Johnson et al., 2015a). For example,  
62 the work on the monitoring of the groundwater/river water interaction beneath the Hanford 300 Area infiltration  
63 bonds (Johnson et al., 2012; Johnson et al., 2015b; Slater et al., 2010; Wallin et al., 2013) shows ERT is well  
64 suited for monitoring such complex and dynamic processes, while the successful monitoring of vadose zone  
65 desiccation (Truex et al., 2013, 2012) at the BC Cribs Area demonstrates its capability to monitor 3-D changes  
66 in moisture content caused by gas injection. The leak tank experiments in the 1990s and 2000s have contributed  
67 some important work in geoelectrical leak detection. The first two mock tank experiments set up a 15 m diameter  
68 steel tank at the Hanford site and ERT tomograms clearly shows area of resistivity decrease of the leak plume  
69 (Ramirez et al., 1996). A subsequent series of mock tank experiments evaluated a number of electrical methods  
70 for leak detection (Barnett et al., 2003). Among them, a “blind test” was carried out for 110 days where the  
71 release episodes were not known to the modeller (Daily et al., 2004). The modeller achieved a 57% success rate  
72 in defining a leak or no leak declaration during the test, although further analysis have greatly improved the

73 success rate. A follow-up study on the dataset used Markov chain Monte Carlo inversion to estimate the  
74 probability distribution of the plume of being in different sizes and shapes (Ramirez et al., 2005).

75 In groundwater hydrology or hydrogeophysical problems, models are often too complex (in terms of  
76 parameterisation) such that fully Bayesian methods such as Markov chain Monte Carlo (MCMC) methods are  
77 rarely applied (Irving and Singha, 2010). Data assimilation has played an increasingly important role in  
78 subsurface characterization (Zhou et al., 2014). For example, Chen et al. (2013) used p-space ensemble Kalman  
79 filter (EnKF) (Nowak, 2009; Schöniger et al., 2012) and ensemble smoother (ES) to assimilate head, flowmeter,  
80 and conservative tracer test data to characterize the permeability field of the Hanford 300 area. Zovi et al. (2017)  
81 used surface ERT results to generate facies model that honour the geophysical data, then used restart normal-  
82 score EnKF to estimate the hydraulic conductivity (K) field. In a recent review, it was concluded that the  
83 iterative ES (IES) could achieve results comparable with those of the EnKF, at a fraction of EnKF's  
84 computational cost (Li et al., 2018). This computational saving stems from the difference in their formulation—  
85 in the EnKF, the data are sequentially integrated into the model at simulation time steps while in ES all the data  
86 are combined together and assimilated only once (note in IES the amount of data between updating steps are  
87 the same). Since EnKF assimilates data in a sequential fashion (i.e. one time step after another), the number of  
88 assimilation steps equals the number of time steps present in the data. Therefore, EnKF is more computationally  
89 expensive than IES when data from many time steps are used.

90 The Hanford leak tank studies and other earlier work on geoelectrical leak monitoring have focused on  
91 obtaining time-lapse ERT images during the suspected leak, and making “leak/ no leak” decisions based on the  
92 images. It is difficult, however, to use geophysical images to infer leak parameters such as leak location, solute  
93 loading, and onset time. Recent hydrogeophysical studies have attempted to estimate parameters of interest from  
94 geophysical data without inverting for geophysical images. Different hydrological model proposals are  
95 evaluated and compared to observed geophysical data. For example, Manoli et al. (2015) used an iterative  
96 particle filter approach and a coupled hydrogeophysical forward model to estimate hydraulic conductivity,  $K$ ,  
97 of up to four zones from ERT data obtained during a controlled infiltration experiment. This approach is then  
98 extended to a field study which considers both ERT and ground penetrating radar (GPR) data in  $K$  estimation  
99 (Rossi et al., 2015). Scholer et al. (2012) used time-lapse crosshole ground GPR data collected under different  
100 infiltration conditions to estimate unsaturated soil hydraulic properties using a MCMC inversion. Kowalsky et  
101 al. (2005) jointly estimated the dielectric and unsaturated zone parameters using both GPR and hydrological  
102 data. Johnson et al. (2009) developed a data-domain correlation approach for joint hydrogeological inversion of  
103 time-lapse hydrogeological and ERT data to jointly estimate fluid solute concentration and resistivity without  
104 explicitly specifying a petrophysical transform.

105 Though contaminant source identification has been a persistent problem in hydrogeology (Michalak  
106 and Shlomi, 2007; Shlomi and Michalak, 2007; Sun, 2007; Sun et al., 2006; Sun and Sun, 2015), advances in  
107 data assimilation methods have opened a new avenue in addressing this problem. Only a few studies have jointly  
108 estimated leak parameters and hydraulic parameters (Datta et al., 2009; Koch and Nowak, 2016; Wagner, 1992).  
109 Zeng et al. (2012) developed a sparse grid Bayesian method for contaminant source identification, which greatly  
110 reduced the computational burden in MCMC sampling and accurately identifies both leak parameters and time-  
111 varying source strengths in case studies. Xu et al. (2016) simultaneously identified the above contaminant source  
112 parameters using the restart normal-score ensemble Kalman filter, while subsequently Xu et al. (2018) extended  
113 the method to also identify the heterogeneous hydraulic conductivity field. The method has recently been  
114 applied to a sandbox study (Chen et al., 2018), where six leak parameters and 2 parameters for the location of  
115 an impermeable plate are estimated. Assuming known source location, Kang et al. (2018) estimated  $K$  and Dense  
116 Non-Aqueous Phase Liquid (DNAPL) saturation (and thus total DNAPL volume) from ERT data using restart  
117 EnKF.

118 In contaminated land studies, there has been a paradigm shift to focus more on site-wide metrics.  
119 Instead of focusing on thresholds from point-based measurements, mass discharge and mass flux has been used  
120 increasingly (Brusseau and Guo, 2014; Christ et al., 2010, 2006; Hadley and Newell, 2012). Several studies are  
121 dedicated to studying their estimation and uncertainty bounds from point measurements (Cai et al., 2011;  
122 Troldborg et al., 2012, 2010), while Balbarini et al. (2018) used regression kriging of collocated concentration  
123 and geoelectrical data to improve mass discharge estimates.

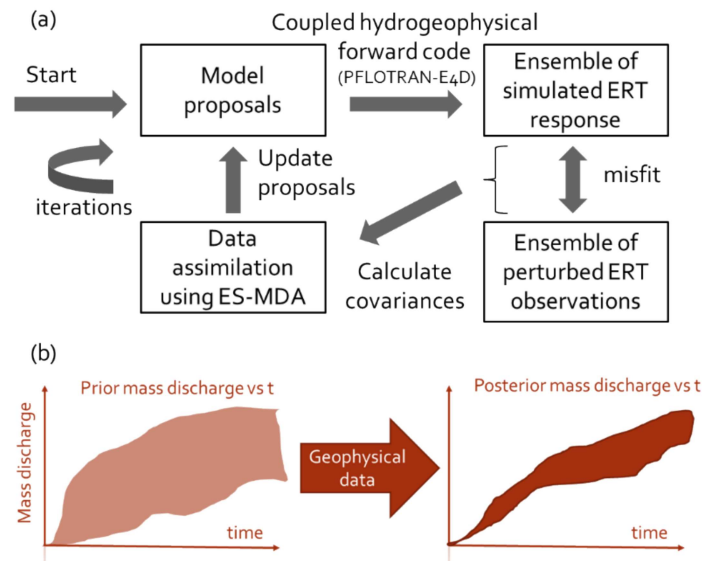
124 In this paper, we introduce an ensemble-based data assimilation framework to jointly identify various  
125 leak parameters with their associated uncertainty bounds from ERT data. The method evaluates proposed  
126 hydrological models (i.e. different hydrogeological units, different leak locations and loads) against observed  
127 time-lapse ERT measurements. To the best of our knowledge, this work is the first attempt to estimate solute  
128 source parameters using raw ERT data, as most previous work focuses on estimating hydraulic parameters or  
129 reconstructing solute distribution. A key feature of our method is that it allows visualization of uncertainty  
130 reduction by comparing the envelopes of prior and posterior mass discharge curves. This method is particularly  
131 suitable for sites where characterization work had been conducted so that previous results can be used to inform  
132 the proposal of prior models. The methods and data used in this work are detailed in section 2. Results of the  
133 various synthetic and field test cases are reported in section 3 and 4 respectively. Finally, we discuss and  
134 summarize our findings in section 5 and 6 respectively.

## 2) Methodology

135 We begin by outlining the different steps in the framework, followed by details of the different  
136 framework components. Finally, we introduce the datasets used in test cases.

### 2.1. Overview of framework

137 The data assimilation framework (summarized by Figure 1(a)) begins by proposing a range of  
138 hydrological models (i.e. model parameters such as leak locations). All parameters for variably saturated flow  
139 and transport simulation need to be prescribed, either as a fixed constant or a distribution (which will be updated  
140 by the DA framework). Also, the setup for the ERT experiment (e.g. mesh, electrode locations, measurement  
141 protocols, petrophysical transforms) need to be included. Once we have an ensemble of model proposals, they  
142 are fed to simulate the ERT response using PFLOTTRAN-E4D (Johnson et al., 2017). The misfits between  
143 observed and simulated ERT responses are used to form data error covariance matrices, which in turn are used  
144 to update the model proposals. The entire process repeats until the misfit criterion is met or the algorithm reaches  
145 the user-specified maximum number of iterations.  
146



147 **Figure 1 (a) Flowchart of the overall data assimilation framework used in this work. More details are found in the subsections.**  
 148 **(b) The goal of this framework is that upon conditioning of geophysical data, the envelope of possible mass discharge time series**  
 149 **will become less uncertain.**

## 150 2.2. Coupled hydrogeophysical forward modelling

151 We use the massively parallel code PFLORAN-E4D (Johnson et al., 2017) for coupled  
 152 hydrogeophysical forward modelling. E4D (Johnson et al., 2010) is an ERT code which has state-of-the-art  
 153 capability for parallelization and for accurate modelling of metallic infrastructure (e.g. tanks and pipes that are  
 154 common at contaminated sites) (Johnson and Wellman, 2015) and near-real-time inversion to monitoring  
 155 bioremediation (Johnson et al., 2015c). E4D has been used for ERT modelling on a number of complex  
 156 problems such as those at the Hanford Site. PFLORAN (Hammond and Lichtner, 2010, also see pfloran.org)  
 157 is a state-of-the-art massively parallel subsurface flow and reactive transport code. PFLORAN-E4D  
 158 (implemented as “hydrogeophysics” mode in the 2018 PFLORAN distributions used in this work) translates  
 159 states of the PFLORAN model to bulk electrical conductivity  $\sigma_b$  distribution using an interpolation matrix that  
 160 maps between the meshes of the two codes given a petrophysical transform. To do so, users need to provide  
 161 elementwise petrophysical parameters (e.g. Archie parameters), times when the simulated ERT measurements  
 162 are needed, and the fluid conductivities of the groundwater and the injected tracer. In this work, we assume  
 163 surface electrical conductivity is negligible and use Archie’s law as the petrophysical relationship:

$$164 \sigma_b = \sigma_w \Phi^m S_w^n \quad (1)$$

165 where  $m$ , is the cementation exponent, and  $n$  is the saturation exponent. Specifically, fluid conductivity  $\sigma_w$ ,  
 166 porosity  $\Phi$ , and fluid saturation  $S_w$  are passed from the PFLORAN output to E4D through the mapping routine.  
 167 After the petrophysical mapping, E4D will run a forward simulation with the given ERT survey configuration  
 168 and  $\sigma_b$  distribution to produce the simulated ERT data. Note that PFLORAN-E4D is no longer supported in  
 169 newer PFLORAN releases. The mapping routine is available through the corresponding author.

## 170 2.3. Prior parameter generation: Latin hypercube sampling

171 For multi-parameter data assimilation problems, we need to use an efficient scheme to generate  $n$  realizations  
 172 model proposals. We use Latin hypercube sampling (LHS) to obtain multi-parameter model proposals that  
 173 efficiently span the parameter space. The LHS approach is implemented using the R package *Envstats* (Millard,  
 174 2013). For the synthetic and field examples, we assume multivariate Gaussian distribution ( $N_e = 32$ ) and

175 multivariate uniform distribution ( $N_e = 64$ ) for the prior distribution of parameter values respectively. The use  
 176 of more realizations and a non-informative prior in the field example is due to greater parameter uncertainty.  
 177

#### 178 **2.4. Data assimilation: ensemble smoother with multiple data assimilation (ES-MDA)**

179 In this work, we use the ensemble smoother with multiple data assimilation (ES-MDA) (Emerick and  
 180 Reynolds, 2013) to update hydrological models. ES-MDA is also known as an iterative variant of ensemble  
 181 smoother (ES). The ES-MDA has been used heavily in hydrocarbon reservoir history matching of production  
 182 and seismic data, but there are growing applications in hydrology. For example, Ju et al. (2018) combined ES-  
 183 MDA with Gaussian process surrogate modelling and tested the new method on synthetic 2-D transient  
 184 groundwater flow problems. Lan et al. (2018) combined sequential ensemble-based optimal design and ES-  
 185 MDA to accurately and efficiently estimate the heterogeneous distribution of physical and geochemical  
 186 parameters in groundwater models. Aalstad et al. (2018) used ES-MDA and fractional snow-covered area  
 187 retrieved from satellites to estimate the snow distribution at Arctic sites. Song et al. (2019) used ES-MDA with  
 188 level set parameterization to estimate the three-facies heterogeneous permeability field at the Hanford IFRC  
 189 site, while Kang et al. (2019) jointly assimilated ERT and concentration data using ES-MDA alongside with  
 190 direct sampling (Mariethoz et al., 2010) to estimate the non-Gaussian hydraulic conductivity field from a  
 191 synthetic salt injection experiment. More recently, a modified version of ES-MDA has been used for crosshole  
 192 GPR travel-time tomography in conjunction with approximate forward solvers and model error correction  
 193 (Köpke et al., 2019).

194 An ensemble smoother (ES) considers all available time-lapse data simultaneously for updating the  
 195 model parameters. The ES-MDA method essentially allows iterative updating of the nonlinear ES problem by  
 196 inflating the observational errors by a factor  $\alpha$  and solve the updating equation  $\alpha$  times iteratively. It has been  
 197 shown that iterative updating better handles nonlinearity in the data assimilation problem than the classic ES  
 198 formulation. Our implementation of the ES-MDA procedure is summarized below:

- 199 1. Prepare observational data (and their error estimates) to be used for data assimilation (DA)
- 200 2. Set up a base PFLOTRAN-E4D model
- 201 3. Decide on which parameter(s) to update, either based on expert judgement or some preliminary global  
 202 sensitivity analysis. The parameter estimation may be affected if important parameters are neither assumed  
 203 correctly nor updated. Sample  $N_e$  realizations from the prior distribution of parameter(s) values (e.g. assume  
 204 normal or uniform distribution) to obtain parameter array  $\mathbf{m}$  at  $l = 0$  ( $\mathbf{m}_0$ ). Parameters that are not being  
 205 updated are assumed known and base model values are used throughout the DA process for all realizations.
- 206 4. Run PFLOTRAN-E4D using  $\mathbf{m}_0$  to obtain an ensemble of simulated ERT data
- 207 5. Updating. For  $l = 1$  to  $N_a$  (where  $N_a$  is the number of data assimilation steps),

- (i.) The data misfit from the  $(l - 1)$ -th iteration is given by

$$209 \quad \text{misfit} = \frac{\sum_{i=1}^{N_d} \sum_{j=1}^{N_e} (d_{obs,j} - d_{i,j})}{N_e \times N_d} \quad (2)$$

210 where  $N_d$  is the number of measurements and  $d_{i,j}$  is the  $j$ -th simulated data of the  $i$ -th realization.

- (ii.) Obtain the auto covariance matrix of model predictions  $\mathbf{C}_{DD}$  and the cross-covariance matrix  
 213 between the parameter vector and model predictions  $\mathbf{C}_{MD}$  by

$$214 \quad \mathbf{C}_{DD} = \text{cov}(\mathbf{d}^i, \mathbf{d}^i) \approx \frac{1}{N_e - 1} \sum_{i=1}^{N_e} (\mathbf{d}_i - \bar{\mathbf{d}})(\mathbf{d}_i - \bar{\mathbf{d}})^T \quad (3)$$

$$215 \quad \mathbf{C}_{MD} = \text{cov}(\mathbf{m}^i, \mathbf{d}^i) \approx \frac{1}{N_e - 1} \sum_{i=1}^{N_e} (\mathbf{m}_i - \bar{\mathbf{m}})(\mathbf{d}_i - \bar{\mathbf{d}})^T \quad (4)$$

where  $\mathbf{d}_i$  and  $\mathbf{m}_i$  are vectors of simulated data and model parameter estimates of the  $i$ -th realization, respectively. The overbar denotes the mean across realizations of a matrix.

- (iii.) For each ensemble member, perturb the observation vector using

$$\mathbf{d}_{uc} = \mathbf{d}_{obs} + \sqrt{\alpha_l} \mathbf{C}_D^{1/2} \mathbf{z}_d \quad (5)$$

where  $\alpha_l$  is an inflation coefficient,  $\mathbf{z}_d \sim N(0, \mathbf{I}_{N_d})$ ,  $\mathbf{I}_{N_d}$  is an identity matrix of size  $N_d$ ,  $\mathbf{C}_D$  is the covariance matrix of the measurements error,  $\mathbf{d}_{obs}$  is a vector of the observed field data. Resampling the vector of perturbed observations at each iteration tends to reduce sampling problems caused by matching outliers that may be generated when perturbing the observations (Emerick and Reynolds, 2013).

- (iv.) Update the parameter ensemble using:

$$\mathbf{m}_l = \mathbf{m}_{l-1} + \underbrace{\mathbf{C}_{MD}(\mathbf{C}_{DD} + \alpha_l \mathbf{C}_D)^{-1}}_{\text{Kalman gain}} (\mathbf{d}_{uc} - \mathbf{d}_{l-1})^T_{\text{misfit}} \quad (6)$$

Note that in order to preserve the equivalence between single and multiple data assimilation, it is necessary that  $\sum_l^{N_a} 1/\alpha_l = 1$  (Emerick and Reynolds, 2013). This effectively serves to update the average sensitivity matrix.

- (v.) Run PFLOTRAN-E4D  $N_e$  times using  $\mathbf{m}_l$  to obtain the updated simulated data ensemble

7. If solution does not converge, repeat steps 3-6 with a higher  $\alpha$  and/or  $N_e$ . Convergence is based on the ensemble root-mean-square-error of the ERT data misfit:

$$RMSE = \sqrt{\frac{1}{N_d N_e} \sum_{i=1}^{N_d} \sum_{j=1}^{N_e} (d_i^{obs} - d_{i,j}^{sim})^2} \quad (7)$$

ES-MDA outperforms ES because the smoother effectively represents a single Gauss–Newton iteration with a full step and an average sensitivity matrix (Reynolds et al., 2006) that is approximated by the covariance matrices of the prior ensemble. Instead of a single and potentially large Gauss-Newton correction, ES-MDA allows multiple smaller corrections through the use of multiple iterations and inflating the covariance matrices to damp the parameter updating (Emerick and Reynolds, 2013). It is more flexible and easier to implement than Gauss-Newton methods because it does not require derivation of sensitivity matrices. Previous work have shown that good results can be obtained in a few iterations (e.g. 4-10), while using a decreasing order of  $\alpha_l$ 's only resulted in small improvements compared to using constant  $\alpha_l$ 's.

In this work, the above steps (except forward modelling) were implemented in R. For the synthetic studies presented, we set  $N_a$  to 7 and use a constant  $\alpha_i$  of 7, which appears to obtain convergence in all cases and also satisfies the criterion  $\sum_l^{N_a} 1/\alpha_l = 1$ . Because the initial misfit for the field data is much larger than that for the synthetic data, the algorithm was unstable and more difficult to converge. Thus, for our field study we set a constant  $\alpha_l$  to 200 and iterate until the RMSE is stabilized, which is achieved within ten iterations. Although this violates the  $\sum_l^{N_a} 1/\alpha_l = 1$  criterion, we remark that its choice is determined based on data noise levels and discrepancy between observed and simulated data, which can be high in field data. A higher  $\alpha_l$  can be seen as adding regularization to the ensemble Kalman scheme (Iglesias, 2016). An alternative approach is to adaptively decide  $\alpha_i$  at each iteration automatically (e.g. Le et al., 2016) based on the mean of RMSE of data misfit across all realizations.

In previous hydrogeology applications using ensemble Kalman methods, the hydraulic heads or solute concentrations are often transformed using normal-score transformation (e.g. Schöniger et al., 2012). We consider ERT data to be more Gaussian than hydrogeological data so we use raw ERT data (transfer resistances)



257 directly in this study but such scaling may improve results. Note that the geometric factors for the crosshole  
258 measurements in our examples do not vary greatly.

259

### 260 **2.5. Plume mass discharge**

261 Mass discharge is the integral of solute fluxes across a control plane (ITRC, 2010). The control plane  
262 can be a model or site boundary, the water table, or any arbitrary planes. Mass flux is defined as  $J = q_0C$ , where  
263 is  $q_0$  groundwater flux and  $C$  is solute concentration. It follows that the solute mass discharge (or equivalently  
264 solute integral flux) across a control plane is defined as  $M_d = \int_A J dA$ , where  $A$  is area of the control plane and  $J$   
265 is the spatially variable solute mass flux. Note that since the solute fluxes are vectors, it is possible for solute  
266 mass discharge to be negative. As shown in Figure 1(b), one way to visualize reduction in site-wide uncertainty  
267 is by observing a reduction of spread of the mass discharge time series.

## 268 **3) Synthetic experiments based on the Sellafield ERT field trial**

269 Between 2013-2014, a field ERT trial was conducted at the Sellafield Nuclear Site in Cumbria, U.K.  
270 (Kuras et al., 2016; Tso et al., 2017) by the British Geological Survey to demonstrate the utility of a permanent  
271 ERT monitoring system to support critical decommissioning activities at nuclear sites. Four vertical boreholes  
272 and two inclined boreholes with forty electrodes each were installed in front of the Sellafield MSSS building.  
273 The field trial included three controlled injections of an electrically conductive tracer (as simulant of the silo  
274 liquor) into the vadose zone. Time-lapse ERT data were collected during the experiment.

275 We built a PFLOTRAN model based on the hydrogeological model developed for Sellafield (Kwong and  
276 Fowler, 2014) and an E4D model based on the electrode locations and design of the field trial. Details of the  
277 PFLOTRAN and E4D models are found in Table 1. Note that there are multiple units in the domain, but only  
278 the hydraulic parameters in the main unit (i.e. sandy drift) is listed in Table 1. The parameters not being  
279 estimated are kept constant during parameter estimation.

280 To test our method, we obtained synthetic ERT data based on the experimental setup of the field trial  
281 and consider a series of parameter estimation cases. They are summarized in Table 2. Unless otherwise stated,  
282 the parameters not being estimated are assumed to be known exactly. We began by considering the estimation  
283 of leak location ( $xloc, yloc$ ), both for a leak inside and outside the ERT monitoring cell. Then we proceed by  
284 also estimating the solute loading ( $q$ ), release onset time ( $t_0$ ). Subsequently, we estimate both leak parameters  
285 and uniform Archie parameters ( $m, n$ ) jointly, which is important in field applications as fixing the parameters  
286 imposes too much confidence on uncertain petrophysical relationships. Finally, we consider a few cases with  
287 uncertainty and heterogeneity in hydraulic conductivity ( $K$ ). In the first case, the  $K$  field has a log variance of  
288 1.0 but its mean value is unknown; while in the second case, the  $K$  field is heterogeneous but its mean value is  
289 known. In the last case, the mean  $K$  value is being estimated for a heterogeneous field. Other potential  
290 parameters to consider includes water table depths, permeability [ $\log_{10}(\text{m}^2)$ ], porosity, unsaturated zone van  
291 Genuchten parameters, recharge rates, depth of the leak ( $zloc$ ), and duration of the leak ( $dt$ ). Each iteration  
292 takes 40 minutes on average to run on 192 cores on PNNL's institutional computing facility. Note that only the  
293 forward modelling is parallelized, not the parameter updating.

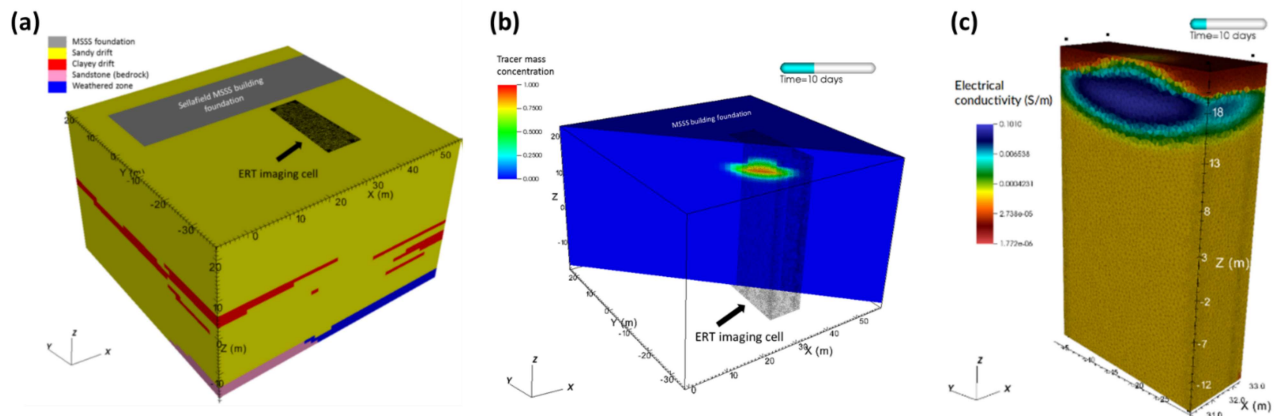
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**Table 1 “True” coupled hydrogeophysical model parameters used for synthetic experiments. It is developed based on the Sellafield field trial. \*Only parameters for the main zone are listed below. #Leak location for some cases is (33.4534, -14.4303) instead. Note that for all cases the leak location is at the water table.**

PFLOTRAN simulation	Value
Total simulation time (days)	30
Model dimensions (m)	40 x 40 x 20
Grid spacing (m)	1 x 1 x 1
Horizontal permeability (m <sup>2</sup> ) *	8.8854 x 10 <sup>-10</sup>
Vertical permeability (m <sup>2</sup> ) *	4.4427 x 10 <sup>-11</sup>
Porosity *	0.2
Water table depth (m)	6.0
van Genuchten <i>m</i>	0.5
van Genuchten $\alpha$	1 x 10 <sup>-4</sup>
Residual water saturation	0.1
Leak location (m) #	(20,-10,18.1)
Leak period (day)	12-30
Leak rate (m <sup>3</sup> /d)	8.0
Background fluid conductivity (S/m)	1 x 10 <sup>-4</sup>
Leak fluid conductivity (S/m)	0.1
Mass discharge plane	Vertical plane at y=-25.03m
E4D simulation	Value
Full Model dimensions (m)	100 x 100 x 100
Imaging cell dimensions (m)	9.5 x 22.8 x 41.5
Grid spacing	Unstructured
Number of elements	380457
ERT imaging times (day)	Every 5 days between day 5 to day 30
Archie's cementation exponent	1.3
Archie's saturation exponent	2.0

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**Figure 2 . (a) PFLOTRAN model domain for the Sellafield MSSS. The grey area is the MSSS building, which is modelled as impermeable. The hashed area is the ERT imaging cell consisting of four ERT boreholes. (b) A snapshot of the simulated tracer concentration due to injection. (c) The corresponding distribution of electrical conductivity within the ERT imaging cell obtained via petrophysical transform.**

304

**Table 2 Summary of synthetic cases. All cases converge in seven iterations.**

Figure	Size of ensemble ( $N_e$ )	Parameter(s) to estimate	Prior distribution	Comments	Initial and final RMSE
Figure 3a	32	$xloc, yloc$	6x6 grid (exclude corners) Uniform spacing X range: -5 – 55m Y range: -33 - -3m	Estimation of the leak location on the x,y plane ; leak is located within the ERT cell	3.63 → 1.01
Figure 3b	32	$xloc, yloc$	6x6 grid (exclude corners) Uniform spacing X range: -5 – 55m Y range: -33 - -3m	Estimation of the leak location on the x,y plane; leak location is outside the ERT cell	7.66 → 1.01
Figure 4	32	$xloc, yloc, q, t_0$	Multivariate uncorrelated truncated Gaussian: $xloc = (mean=25.0, sd=20.0, min=-5.0, max=55.0)$ , $yloc = (mean=-18.0, sd=10.0, min=-33.0, max=-3.0)$ , $q = (mean=15.0, sd=10.0, min=0.0, max=30.0)$ , $t_0 = (mean=15.0, sd=10.0, min=0.0, max=30.0)$	Estimation of the 4 leak parameters	7.01 → 1.01
Figure 5	32	$xloc, yloc, q, t_0, m, n$	Multivariate uncorrelated truncated Gaussian: $xloc = (mean=25.0, sd=20.0, min=-5.0, max=55.0)$ , $yloc = (mean=-18.0, sd=10.0, min=-33.0, max=-3.0)$ , $q = (mean=15.0, sd=10.0, min=0.0, max=30.0)$ , $t_0 = (mean=15.0, sd=10.0, min=0.0, max=30.0)$ , $c = (mean=1.6, sd=0.5, min=0.0, max=2.0)$ , $m = (mean=2.5, sd=0.8, min=0.0, max=3.0)$	Joint estimation of leak parameters and uncertain (homogeneous) petrophysical parameters (Archie's cementation factor and saturation exponent)	22.65 → 1.65
Figure 6a	32	$xloc, yloc, q, t_0$	Multivariate uncorrelated truncated Gaussian: $xloc = (mean=25.0, sd=20.0, min=-5.0, max=55.0)$ , $yloc = (mean=-18.0, sd=10.0, min=-33.0, max=-3.0)$ , $q = (mean=15.0, sd=10.0, min=0.0, max=30.0)$ , $t_0 = (mean=15.0, sd=10.0, min=0.0, max=30.0)$	Leak estimation under the influence of permeability heterogeneity	3.30 → 1.10
Figure 6b	32	$xloc, yloc, q, t_0$	Multivariate uncorrelated truncated Gaussian: $xloc = (mean=25.0, sd=20.0, min=-5.0, max=55.0)$ , $yloc = (mean=-18.0, sd=10.0, min=-33.0, max=-3.0)$ , $q = (mean=15.0, sd=10.0, min=0.0, max=30.0)$ , $t_0 = (mean=15.0, sd=10.0, min=0.0, max=30.0)$	Leak estimation under the influence of and uncertain (homogeneous) permeability	6.47 → 1.31 6.66 → 1.70 6.50 → 1.22 6.53 → 1.41 6.54 → 1.27
Figure 7	32	$xloc, yloc, q, t_0, K$	Multivariate uncorrelated truncated Gaussian: $xloc = (mean=25.0, sd=20.0, min=-5.0, max=55.0)$ , $yloc = (mean=-18.0, sd=10.0, min=-33.0, max=-3.0)$ , $q = (mean=15.0, sd=10.0, min=0.0, max=30.0)$ , $t_0 = (mean=15.0, sd=10.0, min=0.0, max=30.0)$ , $K = (mean=-9.0, sd=sqrt(1.0), min=-11.0, max=-7.0)$	Joint estimation of leak parameters and uncertain (homogeneous) permeability values	3.30 → 1.03

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### 3.1. Base cases

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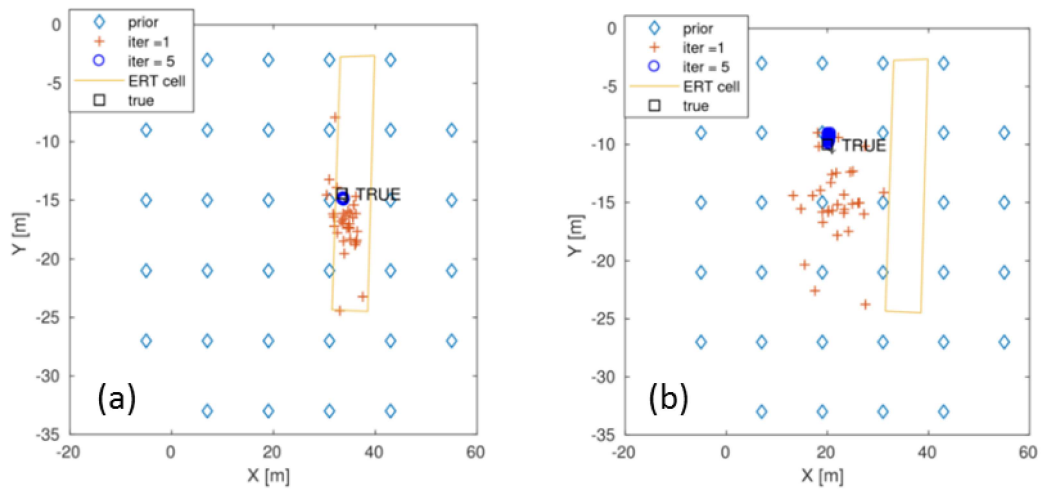
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Our initial example considers the estimation of the leak location (Figure 3). The prior realizations are laid in a rectangular grid. We consider both the cases where the leak is within and outside the ERT imaging cell. Although the estimate at the first iteration is superior when the leak is within the imaging cell, the leak location is accurately estimated after seven iterations in both cases. Figure 4 shows the results from the joint estimation of four leak parameters: the  $(x,y)$  coordinates of the leak location, leak rate, and onset time, assuming a wide multivariate Gaussian prior distribution. After conditioning the parameter values with ERT data, all four leak parameters are accurately estimated. Figure 4b shows the mass discharge curves across a pre-defined plane. The mass discharge curves for the prior distribution are highly variable, while those for posterior distribution collapse to the true curve.



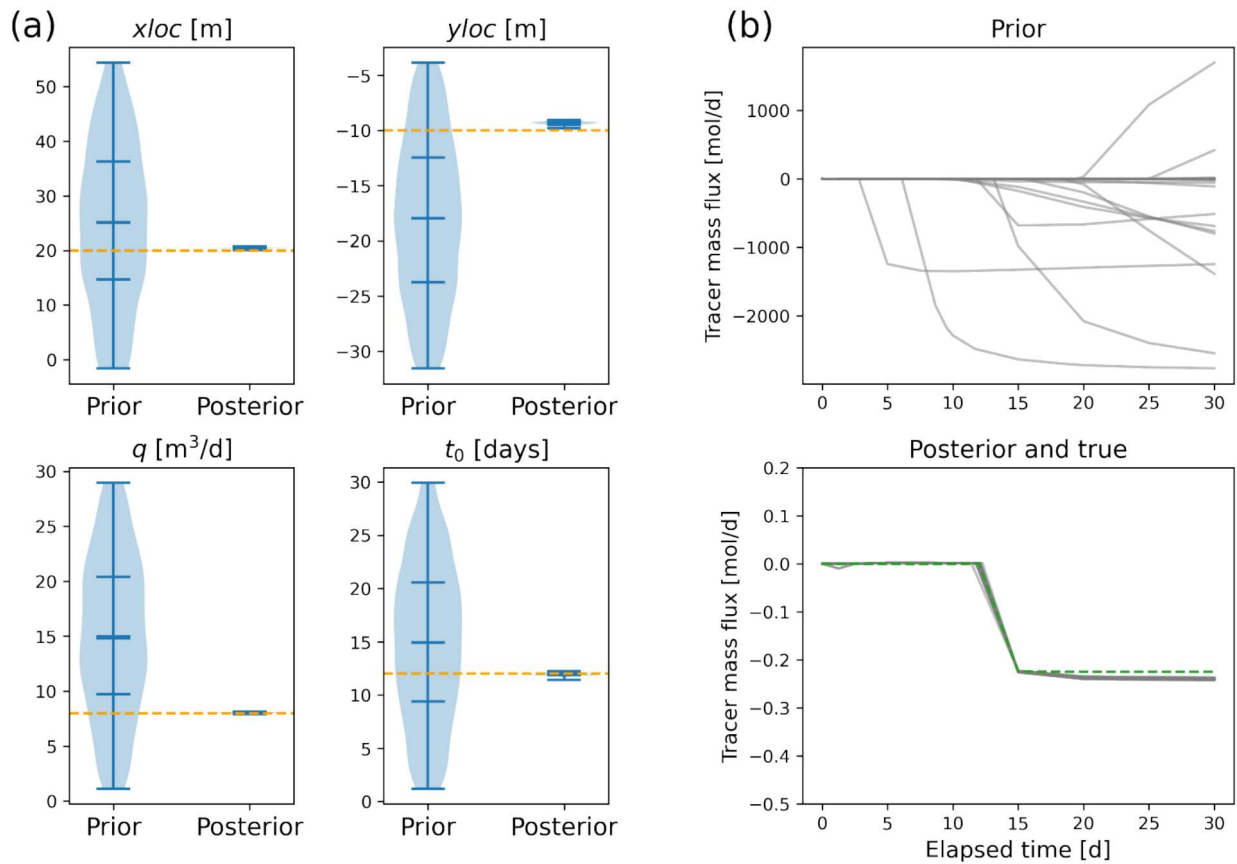
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318 **Figure 3 Estimation of leak location. (a) The true leak location is within the ERT array (33.4534, -14.4303). (b) The true leak**  
319 **location is outside the ERT array (20, -10). In both cases, the data assimilation framework successfully identified the true leak**  
320 **location within a few iterations.**

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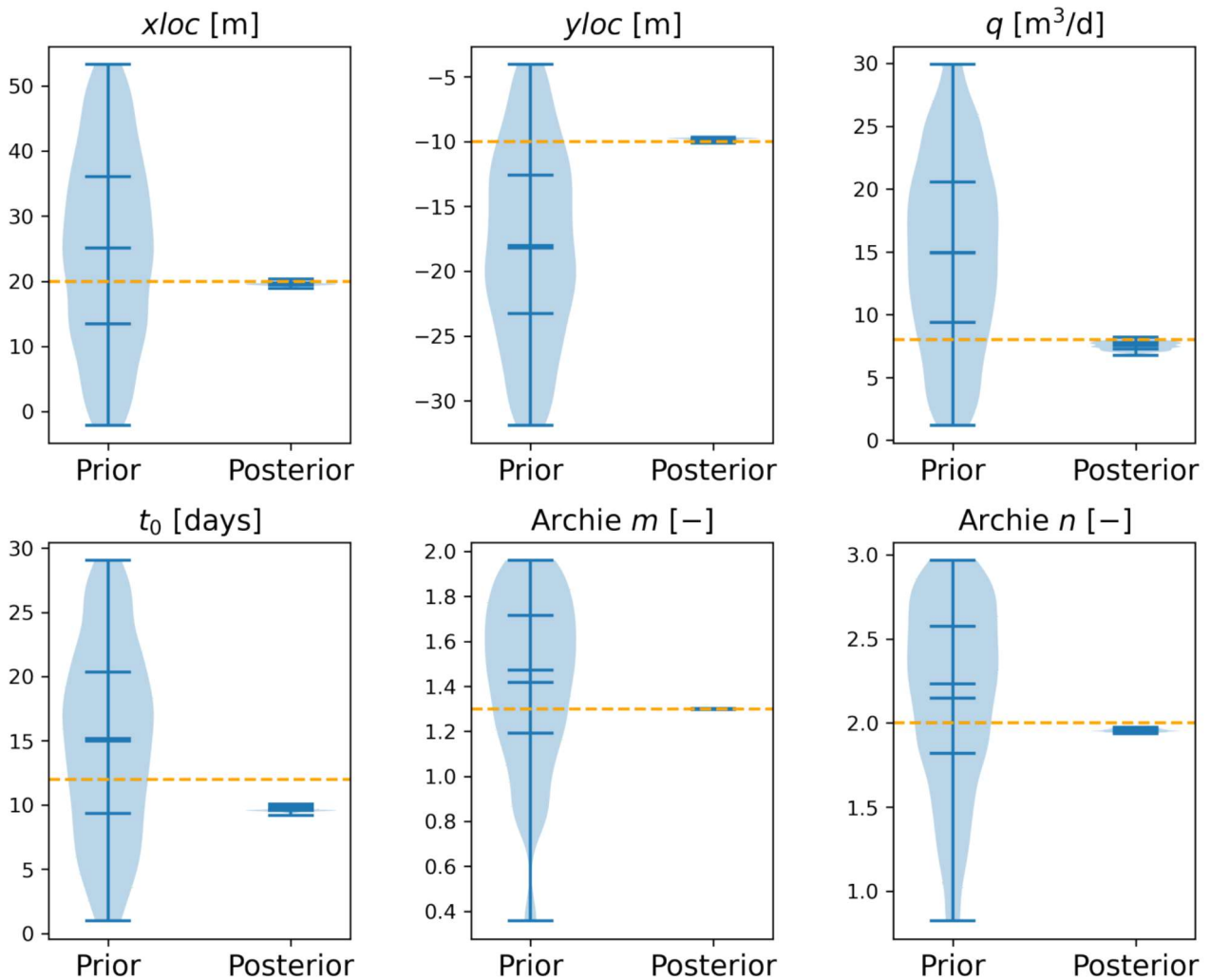
325 **Figure 4** Joint estimation of leak parameters: ( $x,y$ ) location, leak rate, and onset time. (a) Violin plots showing the prior and  
 326 posterior parameter distributions. In each violin plot in this paper, the minima, maxima, mean, median, as well as the 25% and  
 327 75% quantiles are marked. The true values are marked with an orange lines. The posterior parameter values collapse around  
 328 the true values (b) Prior and posterior tracer mass discharge (i.e. integral of mass fluxes) across the pre-defined plane. All the  
 329 posterior curves collapse to nearly the true curve (green). Note that the sign of mass discharge denotes its direction across the  
 330 plane.

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### 332 3.2. Effects of petrophysical parameters

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334 Figure 5 shows the joint estimation of leak parameters and Archie petrophysical parameters. The prior  
 335 estimates are generated as multivariate Gaussian distributions using Latin hypercube sampling. The posterior  
 336 estimates are in very good agreement with the true values, with the exception that the onset time is slightly  
 337 underestimated. It is noteworthy that including the Archie parameters as a covariate has caused the RMSE of  
 338 the prior ensemble to be much higher than those in other synthetic test cases (see Table 2), highlighting that it  
 339 causes a larger range of transfer resistance values.



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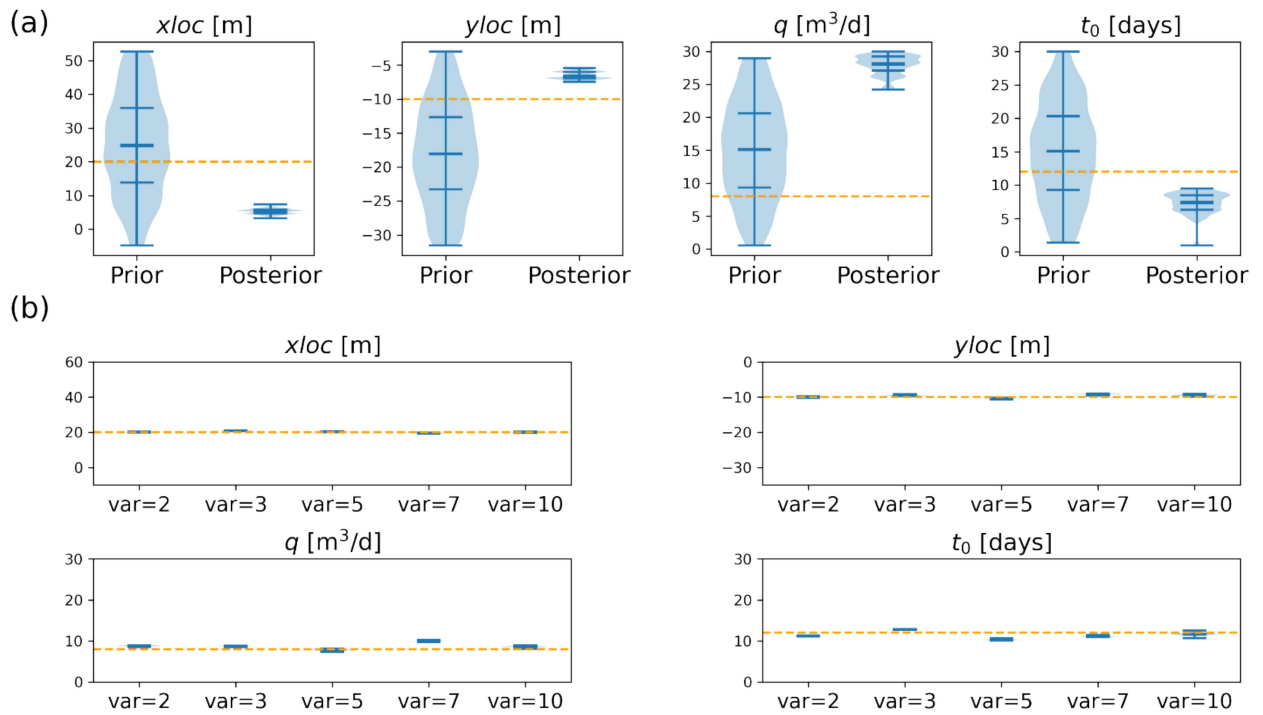
341 **Figure 5** Joint estimation of leak and petrophysical parameters: the prior and posterior parameter distributions are shown as  
 342 violin plots. The true values are marked with orange lines.

343

### 344 **3.3. Influence and joint estimation of uncertain (homogeneous) hydraulic conductivity**

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346 Figure 6a shows the estimation of leak parameters under uncertain  $K$  values. Estimating leak parameters  
 347 under  $K$  uncertainty leads to highly uncertain and inaccurate leak parameter estimates. Figure 6b shows the  
 348 estimation of leak parameters with variance of  $\log K$  equal to 2, 3, 5, 7, 10, while assuming the mean  $K$  values  
 349 are known exactly and unit correlation lengths. Although some variations in the estimates are seen, they  
 350 generally lie close to the true values. There is no apparent correlation between the leak parameter estimation  
 351 performance and the variance of the field. Figure 7 shows the estimates of leak parameters and effective  
 352 hydraulic conductivity. The results show good estimates of the leak locations, while that for  $q$  and  $t_0$  is  
 353 manifested as a narrow envelope. The posterior uncertainty for  $K$  remains high and the algorithm underestimates  
 354 the effective  $K$  value. Again, the envelope of mass discharge curves is greatly reduced, demonstrating a  
 355 reduction in uncertainty. However, the posterior curves do not collapse to the true curve, indicating significant  
 356 uncertainty in the estimates.

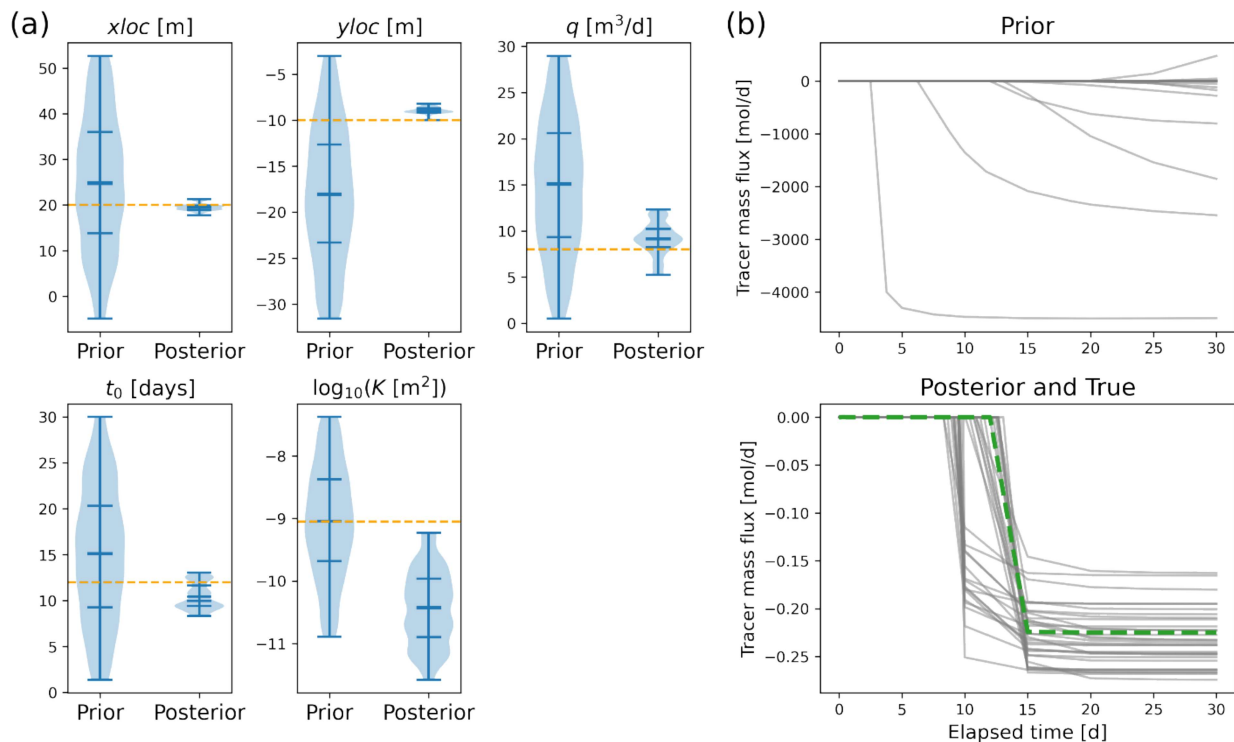


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358 **Figure 6 (a)** The estimation leak parameters under uncertain K values and K ( $\log_{10}$ ) variance = 1.0. The violin plots show the  
 359 prior and posterior parameter distributions. The true value is marked with an orange line. (b) The estimation of leak parameters  
 360 at variance of  $\log_{10}(K)$  equal to 2, 3, 5, 7, 10, while assuming the mean K values are known exactly and the K field is isotropic  
 361 and is of unit correlation length. The violin plots show the posterior parameter distribution, while the true value is marked with  
 362 an orange line.

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Figure 7 (a) Joint estimation of leak parameters and effective hydraulic conductivity. The violin plots show the prior and posterior parameter distributions. The true value is marked with an orange line. (b) Prior and posterior tracer mass discharge across the pre-defined plane. The true curve is marked in green in the posterior plot.

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## 4) Field application at the Hatfield site

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### 4.1. Data description

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To illustrate the approach in a field setting we use data from a solute injection experiment at the Hatfield (Yorkshire) site in the UK. At the site, six boreholes were drilled in 1998 in order to monitor tracer injection, two of which were for transmission GPR measurements (H-R1 and H-R2), while four were for ERT measurements (H-E1, H-E2, H-E3, and H-E4). These four ERT boreholes consist of sixteen stainless steel mesh electrodes equally spaced between 2 and 13 m depth. These boreholes were drilled to a depth of 12 m and completed with 75 mm PVC casing. Both the ERT and radar boreholes have a weak sand/cement grout backfilling the annulus. A tracer injection borehole was also installed (H-I2), located within the centre of the borehole array. The injection borehole is 3.5 m deep, with a 100 mm diameter slotted section and gravel pack between 3 and 3.5 m depth.

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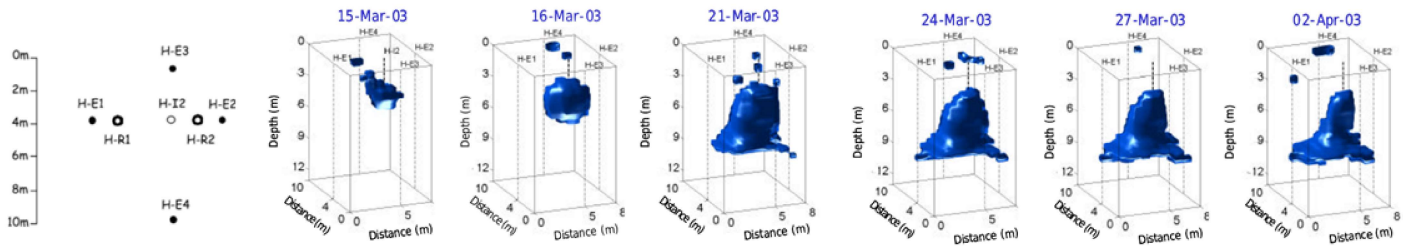
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We focus our discussion using the ERT results from the March 2003 tracer infiltration experiment at Hatfield (Winship et al., 2006). The tracer consisted of 1,200 litres of water, dosed with NaCl to give an electrical conductivity value of  $2200 \mu S cm^{-1}$  (groundwater electrical conductivity at the site was measured as  $650 \mu S cm^{-1}$ ). The tracer was injected over a period of three days, from 14<sup>th</sup> March 2003 to 17<sup>th</sup> March 2003 at a steady rate of approximately 17 litres per hour. A float valve in the injection borehole was used to control the head in the injection borehole, and hence the flow rate. Duplicate sets of background measurements of ERT were made on 6<sup>th</sup> March and 13<sup>th</sup> March. Tracer flow was monitored by means of a pressure transducer in a storage tank, which gave a way of calculating the cumulative injection volume over time. The tracer injection



388 port H-I2 was screened between 3m and 3.5m below ground surface. The tracer injection was monitored by  
 389 ERT measurements from four boreholes and inverted images clearly show the plume migration, as shown in  
 390 Figure 8 (Winship et al., 2006). During the tracer test no rainfall was observed at the site. The water table was  
 391 observed at approximately 10 m depth.

392 After removal of outliers, 3108 of the 3172 measurements are kept and 5% Gaussian data error is assumed  
 393 in the inversions. Let  $t = 8$  be the day where injection commenced, ERT snapshots for  $t = 7, 10, 15, 21$  days  
 394 are used in the inversion. Table 3 lists the baseline parameters for our simulation, which are largely adopted  
 395 from Binley et al. (2002). The parameters not being estimated are kept constant during parameter estimation.



396

397 **Figure 8 Setup of the tracer injection test at Hatfield (H-I2 is the injection borehole and H-E1 to H-E4 are ERT boreholes) and**  
 398 **the time-lapse resistivity images (iso-surfaces are plotted for 7.5% reduction of resistivity relative to baseline) obtained from a**  
 399 **difference inversion of the ERT data (reproduced from Winship et al., 2006)**

400

401 **Table 3 Baseline coupled hydrogeophysical model parameters used for the parameter estimation from the Hatfield field ERT**  
 402 **data. \*The domain consists of 3 meters of top soil and a uniform main zone. Only parameters of the main zone are listed below.**

PFLOTRAN simulation	Value
Total simulation time (days)	41
Model dimensions (m)	30 x 33 x 16
Grid spacing (m)	1 x 1 x 0.5
Permeability (m <sup>2</sup> ) *	4.8225 x 10 <sup>-13</sup>
Porosity *	0.32
Water table depth (m)	-12.0
van Genuchten $m$ *	0.6
van Genuchten $\alpha$ *	3.5 x 10 <sup>-3</sup>
Residual water saturation *	0.04
Recharge (m/day)	1 x 10 <sup>-4</sup>
Leak location (m)	(3.0, 4.0, -3.0)
Leak period (day)	8-11
Leak rate (m <sup>3</sup> /d)	0.408
Background fluid conductivity (S/m)	0.22
Leak fluid conductivity (S/m)	0.065
Mass discharge plane	Vertical plane at $y = -3$ m
E4D simulation	Value
Full Model dimensions (m)	500 x 500 x 50
Imaging cell dimensions (m)	10 x 13 x 15
Grid spacing	Unstructured
Number of elements	46482
ERT imaging times (day) for inversion	7, 10, 15, 21
Archie's cementation exponent	1.35
Archie's saturation exponent	1.35

403

404

405        **4.2. Parameter estimation**

406        We applied the proposed leak detection framework to the Hatfield field data and consider two cases (details  
407        are listed in

408 Table 4). The first case estimates four leak parameters and two Archie parameters ( $xloc, yloc, \log q$   
409 ,  $t_0, m, n$ ). The second case considers the estimation of a few additional parameters, namely the duration of the  
410 leak event ( $dt$ ) and the uniform horizontal and vertical hydraulic conductivity ( $K$  and  $Kz$ ). We consider  $K$   
411 anisotropy may exist at the site because well logs suggest the presence of fine-textured layers (Binley et al.,  
412 2001). Compared to the earlier synthetic examples, convergence was much more difficult to achieve. We have  
413 used the following modification to our methods to circumvent this issue: we estimated  $\log q$  instead of  $q$ , used  
414 more realizations, and used a uniform prior instead of a Gaussian one. We transformed the leak location priors  
415 to a uniform grid to aid the interpretation of the results. We have not considered the estimation of depth of the  
416 leak source in any of our examples because for most leak detection problems, the leak depth is usually precisely  
417 known: for example, base of storage tanks/silos, depth of buried pipelines, and bottom of landfill lining. Each  
418 iteration takes 2.5 hours on average to run on 192 cores. Note that only the forward modelling is parallelized,  
419 not the parameter updating.

420 Results from the base case is reported in Figure 9. Figure 9(a) shows that the posterior estimates of most  
421 parameter pairs form a small cluster. The estimates of  $xloc$  and  $\log q$  are close to the true values, while those  
422 for  $yloc$  and  $t_0$  are slightly above the true (known) values. The inversion appears to have no sensitivity to  $m$ ,  
423 while the estimation of  $n$  converges to a very small value of about 0.53. Note that in this field test the true values  
424 of  $m$  and  $n$  are not known. In the inversion of field data, we would not necessarily consider the estimates of  $m$   
425 and  $n$  representative of actual petrophysical parameters, but rather they act as hyperparameters to adjust any  
426 discrepancy in model structure. Figure 9(b) shows that the variability of mass discharge curves between  
427 realizations is greatly reduced upon conditioning of ERT data. Specifically, its spread is reduced by two orders  
428 of magnitude, highlighting a reduction in site-wide uncertainty of the plume migration.

429 Results from the second case are reported in Figure 10. We observe a larger spread in the parameter  
430 space but similar results for the estimation of  $m$  and  $n$ .  $xloc, yloc$ , and  $t_0$  are slightly overestimated. The  
431 inversion appears to have no sensitivity to  $K$  and  $Kz$ . The estimates of  $\log q$  and  $dt$  centres around the true  
432 value, indicating the inversion algorithm also correctly estimates the total solute loading ( $q \times dt$ ) that enters  
433 the flow and transport modelling domain. This underscores that the proposed data assimilation framework does  
434 not suffer from mass balance issues that are common in inverted resistivity-based approaches.

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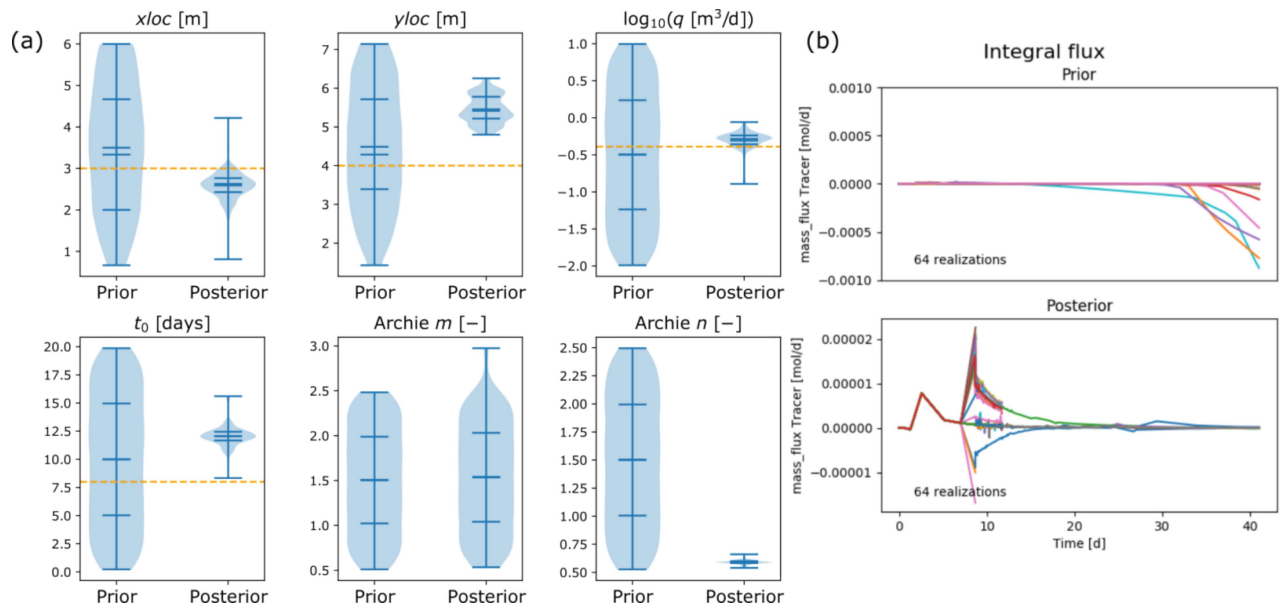
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**Table 4 Summary of cases for the Hatfield field example**

Figure	Size of ensemble ( $N_e$ )	Parameter(s) to estimate	Prior distribution	Comments	Final RMSE
Figure 9	64	$xloc$ , $yloc$ , $\log q$ , $t_0$ , $m$ , $n$	Multivariate uncorrelated uniform: Adjusted uniform grid from $xloc$ 0-8m and $yloc$ =0-10m $\log q = \text{list}(\text{min}=-2.0, \text{max}=1.0)$ , $t_0 = \text{list}(\text{min}=0.0, \text{max}=20.0)$ , $m = \text{list}(\text{min}=0.5, \text{max}=2.5)$ , $n = \text{list}(\text{min}=0.5, \text{max}=2.5)$	Base case	223.16 $\rightarrow$ 15.3 (iter=8, stabilized afterwards)
Figure 10	64	$xloc$ , $yloc$ , $\log q$ , $t_0$ , $dt$ , $m$ , $n$ , $K$ , $Kz$	Multivariate uncorrelated uniform: Adjusted uniform grid from $xloc$ 0-8m and $yloc$ =0-10m $\log q = \text{list}(\text{min}=-2.0, \text{max}=1.0)$ , $t_0 = \text{list}(\text{min}=0.0, \text{max}=20.0)$ , $dt = \text{list}(\text{min}=1.0, \text{max}=5.0)$ , $m = \text{list}(\text{min}=0.5, \text{max}=2.5)$ , $n = \text{list}(\text{min}=0.5, \text{max}=2.5)$ , $K = \text{list}(\text{min}=-13.0, \text{max}=-9.0)$ , $Kz = \text{list}(\text{min}=-13.0, \text{max}=-9.0)$	$K$ , $Kz$ , and $dt$ are also estimated.	310.66 $\rightarrow$ 13.95 (iter = 2, stabilized afterwards)

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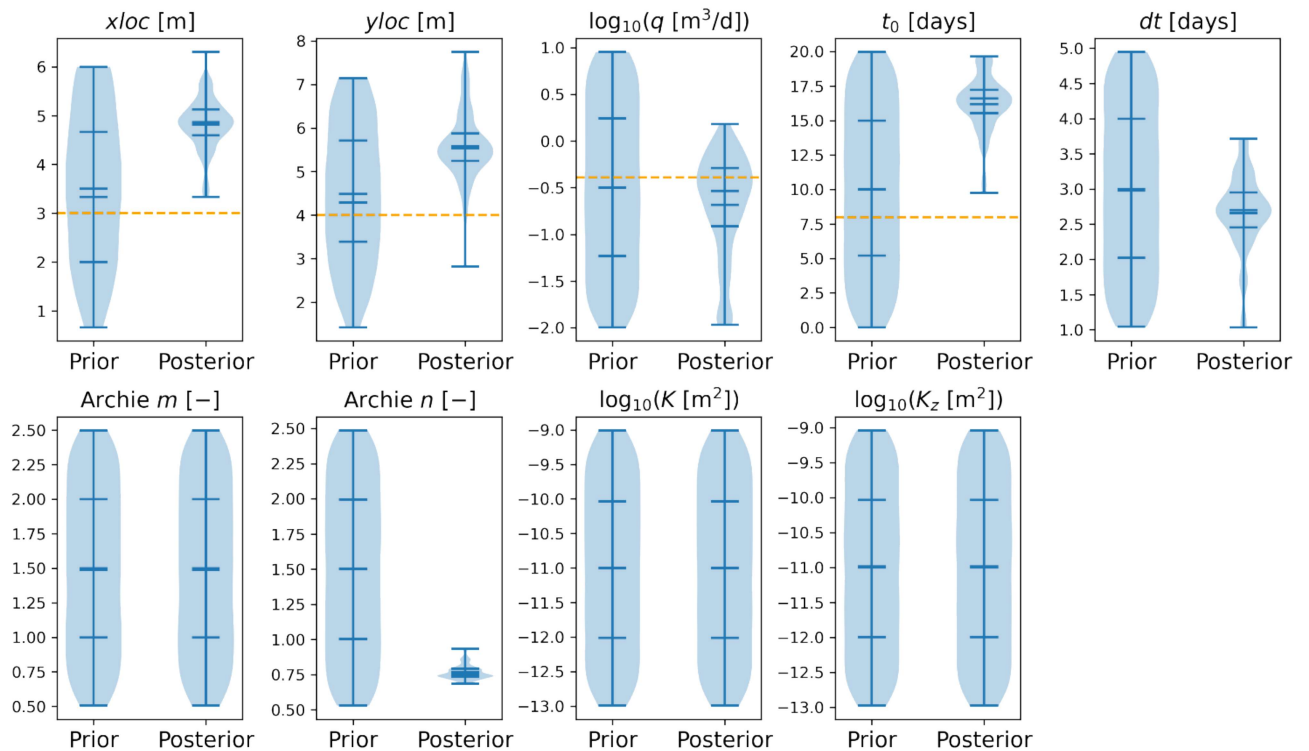
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**Figure 9 (a) Violin plots showing the prior and posterior parameter values for the Hatfield example estimating leak and Archie parameters. The parameter symbols are defined in section 3. The true leak parameters used in the field injection experiment is indicated by the orange lines. (b) The prior and posterior mass discharge time series. The sign of mass discharge indicates the direction across the defined plane.**



444

445 **Figure 10** Violin plots showing the prior and posterior parameter values for the Hatfield example estimating leak and Archie  
 446 parameters and hydraulic conductivities. The parameter symbols and units are defined in Table 3. The true leak parameters  
 447 used in the field injection experiment is indicated by orange lines.

448

### 449 **4.3. Global sensitivity analysis using the Morris method**

450 To better understand the sensitivity of ERT data to various parameters in the coupled hydrogeophysical  
 451 problem used to analyse the Hatfield dataset, we performed a global sensitivity analysis using the Morris  
 452 method (Morris, 1991; Tran et al., 2016; Wainwright et al., 2014) that is implemented in the R package  
 453 *sensitivity* (Iooss, 2019). The Morris method returns the elementary effect (EE) of the parameters, which can  
 454 be considered as an extension of the local sensitivity method. Since the mean EE represents the average effect  
 455 of each parameter over the parameter space, the mean EE can be regarded as a global sensitivity measure. To  
 456 ignore the effects of the sign, the mean of absolute EE is usually reported (mean |EE|). In general, for the  
 457 parameter ranges considered, parameters with high mean |EE| have a large impact on the data. Unconditional  
 458 realizations are generated using the Morris algorithm based on the parameter ranges specified in Table 5 and  
 459 the parameter space is sampled uniformly. We used 25 chains, so for a 13 parameter problem  
 460  $25 \times (13 + 1) = 350$  realizations are generated. We run the forward models using PFLOTRAN-E4D to  
 461 obtain simulated ERT response (the settings are the same as those in

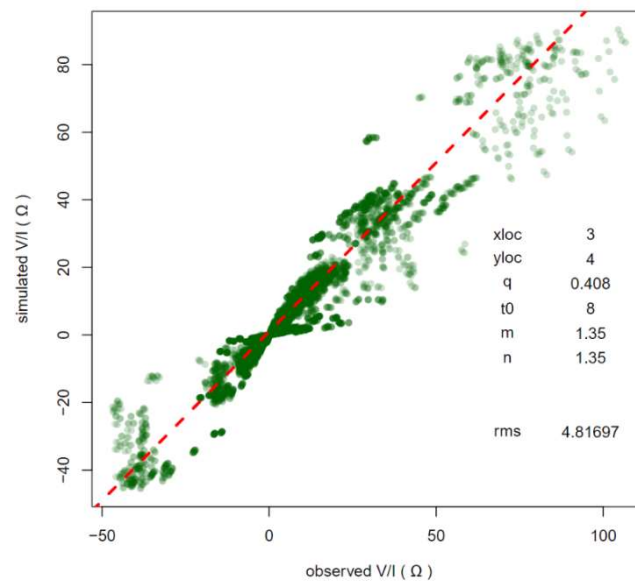
462 Table 4, unless otherwise stated). We set the objective function for calculating the mean  $|EE|$  to be the  
 463 weighted misfit between the simulated and observed ERT data at Hatfield. The same dataset as in the previous  
 464 section is used.

465 Results from the global sensitivity analysis of the Hatfield experiment shows that some parameters,  
 466 especially water table depths and two of the van Genuchten parameters have the largest effects on the data misfit  
 467 (Table 5), followed by uniform permeability, porosity and Archie parameter values. Leak parameters has low  
 468 mean  $|EE|$ , indicating the difficulty for ERT data to inform their estimation if the others are not known with  
 469 confidence. Among them,  $xloc$  and  $yloc$  have the highest and the lowest mean $|EE|$ , respectively. Recharge has  
 470 virtually no effect on the data misfit. The results show that using ERT data and coupled hydrogeophysical  
 471 modelling is a challenging problem. Future work can benefit from better constraining the problem incorporating  
 472 additional data sources (e.g. pressure head, concentration, temperature, saturated and unsaturated hydraulic  
 473 parameters). Our results agree with that of [Tran et al. \(2016\)](#), who showed Archie parameters have a higher  
 474 mean  $|EE|$  than van Genuchten  $\alpha$ . However, they found the mean  $|EE|$  of van Genuchten  $m$  is negligible, while  
 475 the largest mean  $|EE|$  for ERT data they found is around 8.0. This highlights the Morris sensitivity analysis is  
 476 best considered on a case-by-case basis, as it is affected by the observed data and the selected parameter ranges.  
 477 We also report a list of realizations with low data misfit in Table 5. We observe that none of the realizations  
 478 have an RMSE lower than 7.4 and their parameter values vary greatly. It is noteworthy that a “true” deterministic  
 479 run (using parameters in Table 3) would give an RMSE of 4.82 (Figure 11). The above shows that some  
 480 solutions to the ERT leak detection problem can be considered equifinal.

481 **Table 5 Global sensitivity analysis results using the Morris (1991) method on selected parameters on the Hatfield coupled**  
 482 **hydrogeophysical model. The parameter ranges considered and the mean absolute elementary effect ( $|EE|$ ) are reported.**  
 483 **Parameter value combinations from ten realizations with the lowest RMSE are also reported.**

Parameters [units]	Range	Mean $ EE $	#24	#59	#61	#62	#63	#133	#150	#152	#153	#154
$xloc$ [m]	0.0 – 8.0	7.65	0.0	8.0	2.0	2.0	2.0	6.0	2.0	2.0	2.0	2.0
$yloc$ [m]	0.0 – 10.0	0.18	0.0	10.0	2.5	2.5	2.5	10.0	7.5	7.5	7.5	7.5
$q$ [ $\log_{10}$ ( $m^3/d$ )]	-2.0 – 1.0	1.82	-2.00	-1.25	-1.25	-1.25	-1.25	1.00	-1.25	-1.25	-1.25	-1.25
$t_0$ [d]	0.0 – 20.0	1.53	15.00	5.00	5.00	5.00	5.00	15.00	20.00	20.00	20.00	20.00
Archie $m$ [-]	1.0 – 1.5	26.52	1.38	1.00	1.00	1.00	1.00	1.50	1.50	1.50	1.50	1.50
Archie $n$ [-]	0.5 – 2.0	11.68	1.63	0.88	0.88	0.88	0.88	2.00	1.63	1.63	1.63	1.63
water table [m]	-14.0 – -9.0	49.39	-14.00	-14.00	-14.00	-14.00	-14.00	-14.00	-12.75	-12.75	-12.75	-12.75
permeability [ $\log_{10}$ ( $m^2$ )]	-15.0 – -12.0	6.85	-15.00	-12.00	-12.00	-14.25	-14.25	-12.00	-12.00	-12.00	-12.00	-12.00
porosity [-]	0.25 – 0.35	12.34	0.25	0.35	0.35	0.35	0.35	0.28	0.35	0.35	0.28	0.28
VG $\alpha$ [ $Pa^{-1}$ ]	2e-4 – 2e-3	7.50	2.0e-3	2.0e-4	2.0e-4	2.0e-4	1.55e-3	2.0e-4	6.5e-4	6.5e-4	6.5e-4	6.5e-4
VG $m$ [-]	0.4 – 0.8	115.16	0.7	0.7	0.7	0.7	0.7	0.5	0.4	0.7	0.7	0.7
VG $Sr$ [-]	0.01 – 0.2	69.30	0.2	0.01	0.01	0.01	0.01	0.1525	0.01	0.1525	0.1525	0.1525
recharge [mm/d]	0.0 – 0.001	0.03	0.00	7.5e-4	7.5e-4	7.5e-4	7.5e-4	0.00	1.0e-3	1.0e-3	1.0e-3	2.5e-4
RMSE	--	--	7.54	11.10	11.22	11.15	10.57	9.30	8.25	9.40	7.44	7.42

484



485

486 **Figure 11** Transfer resistance scatter plot between the observed and simulated data at Hatfield. The simulated data uses  
 487 parameter values listed in Table 3.

## 488 5) Discussion

489 ERT has been used to detect leaks from nuclear sites for more than two decades. The conventional  
 490 approach is to use inversion to obtain smoothed images of resistivity at different times and to assess whether  
 491 there is a leak. This approach does not allow estimation of leak parameters and inversion of large time-lapse  
 492 ERT datasets can be computationally demanding. We have presented a data assimilation framework to estimate  
 493 leak parameters from ERT data. It evaluates hydrological model proposals based on the misfits between  
 494 simulated and observed ERT data and update the model proposals. The estimated leak parameters are presented  
 495 as a posterior distribution. It also outputs plume mass discharge across a plane, which can be used as a metric  
 496 to evaluate site-wide uncertainty reduction. These features are not available in existing methods. Since current  
 497 methods to estimate mass discharge are based on interpolation of point measurements, our coupled modelling  
 498 approach provides an alternative to quantify mass discharge estimates. Together with point measurements and  
 499 ERT imaging, the various methods can help establish multiple lines of evidence to better inform decision making  
 500 in nuclear site characterisation.

501 Our synthetic results show that the method allows very good estimation of leak parameters (e.g. leak  
 502 rate, loading size, and location). They also show that this framework can work reasonably well under the  
 503 influence of uncertain petrophysical parameters and mean  $K$  values, as well as under  $K$  heterogeneity with small  
 504 correlation lengths (i.e. with no significant structures). With the rapid growth of autonomous ERT systems to  
 505 monitor infrastructure, such as British Geological Survey's ALERT and PRIME (Huntley et al., 2019) systems,  
 506 our approach can provide additional value to ERT data and supplement inverted resistivity images. Our work  
 507 also has potential to be applied to other non-point source leak detection problems such as seepage through  
 508 embankments, or using a different geophysical method such as self potential (SP). Likewise, our method could  
 509 also be applied to induced polarisation (IP) data, which has been shown to be potentially effective for monitoring  
 510 some reactive plume processes.

511 We have only examined problems with a few parameters (e.g. leak parameters and homogeneous Archie  
 512 and permeability values). All hydrological and petrophysical parameters that are not being updated are treated  
 513 as known constants, which can be strong assumptions on uncertain subsurface properties. Future work should

514 strive to relax such assumption and jointly estimate more parameters. The prior distribution of the uncertain  
515 parameters may have an effect on the performance of our data assimilation approach. Nonetheless, we  
516 emphasize that they should be chosen based on site-specific prior knowledge. In this work, we have considered  
517 a relatively simple problem: a single conservative source with known concentration (thus fluid conductivity)  
518 with a single release episode. With the aid of relevant auxiliary information, our framework has the potential to  
519 be extended to more complex problems.

520 The challenges we have encountered when dealing with field data highlights the need of unbiased and  
521 reliable prior information for the proposed method to work in practice. Equifinality (Beven, 2006; Binley and  
522 Beven, 2003) obviously exist in the leak detection problem since multiple combinations of leak, petrophysical,  
523 and hydraulic parameters can give similar data misfits. Different parameterization, scaling of parameters, and  
524 additional data sources may alleviate the problem. But ultimately, methods that allow rejection of model  
525 proposals may be desirable. Nevertheless, our method can be considered both a quick and approximate method  
526 for quantifying posterior uncertainty of parameters of interest, as well as a flexible method to perform  
527 regularized inversion without forming the Jacobian (Iglesias, 2016), which can be advantageous for coupled  
528 problems. Our proposed method is best used in well characterized sites where an abundance of historical data  
529 can be used to build prior models. Alternatively, our method can also be used in controlled tracer injection  
530 experiments to estimate hydraulic, petrophysical and transport parameters.

531 There exists unique challenges for using raw ERT data in data assimilation. ERT datasets are usually  
532 quite large, with each timeframe containing hundreds to tens of thousands of data points. The fast collection of  
533 ERT data mean that multiple datasets can be collected daily. However, due to computation constraints, we have  
534 only used data from a few selected days. Also, each ERT quadrupole measurement neither represent the state  
535 response at a point (as in head or concentration data) or the overall system response (as in hydrocarbon  
536 production rates). These challenges do not appear to impact leak estimation from synthetic results. But their  
537 implications warrant further investigation—for example, can we compress raw ERT data for data assimilation  
538 since they may contain significant redundant information?

539 Frameworks for efficient high-dimensional data assimilation (Ghorbanidehno et al., 2015; Li et al.,  
540 2015, 2016) can be used to jointly estimate a heterogeneous permeability field. Methods such as level set  
541 methods, discrete cosine transform (DCT) and principal component analysis (PCA) can reduce the number of  
542 parameters to describe a highly heterogeneous field. A recent study has applied ES-MDA in combination with  
543 level set methods (Iglesias and McLaughlin, 2011; Tai and Chan, 2004) to estimate the three-facies  
544 heterogeneous permeability field from conservative tracer test data at the Hanford IFRC site (Song et al., 2019).  
545 Future work should explore their utility in hydrogeophysical data assimilation. Likewise, we have assumed  
546 relatively simple petrophysical relationships in our coupled hydrogeophysical models. Whether more complex  
547 petrophysical models will improve data assimilation results remains an open question. We also have not  
548 examined joint assimilation of ERT data with head or concentration data, which can be promising for further  
549 constraining our results. In this paper, we have used a relatively small ensemble of highly detailed, fully coupled  
550 hydrogeophysical simulations as the forward model. Our work can benefit from a recently developed, adaptive  
551 multi-fidelity version of ES-MDA (Zheng et al., 2018), which leverages both the accuracy of highly detailed  
552 models and the efficiency of simplified models within the ES-MDA framework.

## 553 6) Conclusions

554 We propose a data assimilation framework that allows the use of time-lapse ERT data for solving  
555 hydrological parameters in a leak detection problem. It does not produce any ERT images during inversion;  
556 rather, it updates parameters in the hydrological model to minimize ERT data misfit. The use of an ensemble-



557 based framework allows straightforward computation of uncertainty estimates. Site-wide uncertainty reduction  
 558 can be visualized by comparison of prior and posterior mass discharge curves. Synthetic and field results  
 559 demonstrate its utility under a variety of settings, e.g. when uniform hydrological and Archie parameters are  
 560 estimated jointly. This new framework is particularly attractive to sites that have previously undergone  
 561 extensive geological investigation (e.g., nuclear sites). It can be readily extended to solving other complex  
 562 problems (e.g. multiple modalities) of interest that is monitored by geophysical data. We have only used ERT  
 563 data in our analysis but the framework is highly flexible that it is straightforward to incorporate multiple data  
 564 types. Our method complements electrical resistivity imaging and is particularly applicable to sites where some  
 565 prior characterization is performed and uncertainty estimates for the parameters that drive the underlying  
 566 processes observed are desired.

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