



Deconvolution of  
complex atmospheric  
datasets

K. P. Wyche et al.

# Mapping gas-phase organic reactivity and concomitant secondary organic aerosol formation: chemometric dimension reduction techniques for the deconvolution of complex atmospheric datasets

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Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



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**Deconvolution of  
complex atmospheric  
datasets**

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



## Abstract

Highly non-linear dynamical systems, such as those found in atmospheric chemistry, necessitate hierarchical approaches to both experiment and modeling in order, ultimately, to identify and achieve fundamental process-understanding in the full open system. Atmospheric simulation chambers comprise an intermediate in complexity, between a classical laboratory experiment and the full, ambient system. As such, they can generate large volumes of difficult-to-interpret data. Here we describe and implement a chemometric dimension reduction methodology for the deconvolution and interpretation of complex gas- and particle-phase composition spectra. The methodology comprises principal component analysis (PCA), hierarchical cluster analysis (HCA) and positive least squares-discriminant analysis (PLS-DA). These methods are, for the first time, applied to simultaneous gas- and particle-phase composition data obtained from a comprehensive series of environmental simulation chamber experiments focused on biogenic volatile organic compound (BVOC) photooxidation and associated secondary organic aerosol (SOA) formation. We primarily investigated the biogenic SOA precursors isoprene,  $\alpha$ -pinene, limonene, myrcene, linalool and  $\beta$ -caryophyllene. The chemometric analysis is used to classify the oxidation systems and resultant SOA according to the controlling chemistry and the products formed. Furthermore, a holistic view of results across both the gas- and particle-phases shows the different SOA formation chemistry, initiating in the gas-phase, proceeding to govern the differences between the various BVOC SOA compositions. The results obtained are used to describe the particle composition in the context of the oxidized gas-phase matrix. An extension of the technique, which incorporates into the statistical models data from anthropogenic (i.e. toluene) oxidation and “more realistic” plant mesocosm systems, demonstrates that such an ensemble of chemometric mapping has the potential to be used for the classification of more complex spectra of unknown origin. The potential to extend the methodology to the analysis of ambient air is discussed using results obtained from a zero-dimensional box model incorporating mechanistic data obtained from the Master

## Deconvolution of complex atmospheric datasets

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



Chemical Mechanism (MCMv3.2). Such an extension to analysing ambient air would prove a powerful asset in assisting with the identification of SOA sources and the elucidation of the underlying chemical mechanisms involved.

## 1 Introduction

Biogenic Volatile Organic Compounds (BVOCs) are ubiquitous in the global troposphere, being emitted primarily from terrestrial plant life (Kanakidou et al., 2005). It is estimated that the total annual emission rate of all (non-methane) BVOCs is roughly ten times that of all anthropogenic volatile organic compounds, being around 750 Tg C yr<sup>-1</sup> (Sindelarova et al., 2014). With the exception of methane, the most dominant species of BVOCs in terms of emission strength, reactivity and their impact upon the atmosphere, are terpenes (Reinigg et al., 2008) a subdivision of BVOCs that primarily comprise the hemiterpene, isoprene (C<sub>5</sub>), monoterpenes (C<sub>10</sub>) and sesquiterpenes (C<sub>15</sub>) (e.g. Atkinson and Arey, 2003a; Kanakidou et al., 2005).

Within the troposphere terpenes are able to react with OH, O<sub>3</sub> and NO<sub>3</sub> at appreciable rates (e.g. Calvert et al., 2000; Koch et al., 2000; Fantechi et al., 2002; Capouet et al., 2004; Kroll et al., 2006) such that their atmospheric lifetimes are in the order of minutes – hours (e.g. Calogirou et al., 1999). Because of their large emission rates and high reactivities, terpenes have a strong impact upon the chemistry of the troposphere at the local, regional and global scales (e.g. Jaoui and Kamens, 2001; Paulot et al., 2012; Surratt, 2013). For instance, terpenes have high photochemical ozone creation potentials (Derwent et al., 2007) and extensive photochemical oxidation pathways that lead to the production of a complex array of oxygenated and nitrated products, some of which are able to form secondary organic aerosol (SOA) (e.g. Calvert et al., 2000; Capouet et al., 2004; Jenkin, 2004; Baltensperger et al., 2008; Kanakidou et al., 2005; Surratt et al., 2006; Kroll and Seinfeld, 2008; Hallquist et al., 2009).

Aerosol particles are natural components of the Earth's atmosphere responsible for a range of well-documented impacts, ranging from visibility impairment on the local

## Deconvolution of complex atmospheric datasets

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion





data that often only a fraction is subsequently used in a given analysis. For example, during a typical six-hour environmental simulation chamber experiment, VOC monitoring chemical ionisation reaction time-of-flight mass spectrometry, will produce roughly  $1.1 \times 10^7$  data points. In order to keep pace with instrument development and maximise the information extracted from sometimes-complex experiments, it is crucial that we advance our data analysis methods and introduce new data mining techniques.

The work reported here focuses on detailed organic gas-phase and particle-phase composition data, recorded during SOA atmospheric simulation chamber experiments, using chemical ionisation reaction time-of-flight mass spectrometry (CIR-TOF-MS) and liquid chromatography-ion trap mass spectrometry (LC-MS/MS), respectively, as well as broad (i.e. generic composition “type”; oxygenated organic aerosol, nitrated, sulphated etc) aerosol composition data, recorded by compact time-of-flight aerosol mass spectrometry (cTOF-AMS). The goal of this paper is to demonstrate and evaluate the application of an ensemble reductive chemometric methodology for these comprehensive oxidation chamber datasets, to be used as a model framework to map chemical reactivity from mesocosm systems, thus providing a link from model systems to more “real” mixtures of organics. The intermediate complexity offered by simulation chamber experiments makes them an ideal test-bed for the methodology. Application of the methodology to resultant particle-phase data also aims to provide a level of particle composition classification in the context of gas-phase oxidation. Similar approaches using statistical analyses have been recently applied to both detailed and broad ambient aerosol composition data (Heringa et al., 2012; Paglione et al., 2014), particularly in the context of source apportionment (Alier et al., 2013). However our approach investigates both the gas- and particle-phases and also provides insight into the fundamental chemical reaction pathways.

The central methodology employed, is based around the application of *principal component analysis* (PCA), *hierarchical cluster analysis* (HCA) and *positive least squares-discriminate analysis* (PLS-DA) of single-precursor oxidant chemistry in environmental simulation chambers. Colloquially, we can describe these three approaches as provid-

## Deconvolution of complex atmospheric datasets

K. P. Wyche et al.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

ing dimensions along which the data are separable (PCA), tests of relatedness (HCA) and checks for false-positives (PLS-DA).

Such dimension reduction techniques can be very powerful when used in chemometrics, enabling large and often complex datasets to be rendered down to a relatively small set of pattern-vectors to provide an optimal description of the variance of the data (Jackson, 1980; Sousa et al., 2013; Kuppusami et al., 2014).

The analysis conducted shows that “model” biogenic oxidative systems can be clearly separated and classified according to their gaseous oxidation products, i.e. isoprene from  $\beta$ -caryophyllene from non-cyclic monoterpenes and cyclic monoterpenes. The addition of equivalent mesocosm data from fig and birch tree experiments shows that large isoprene and large monoterpene emitting sources, respectively, can be mapped onto the statistical model structure and their positional vectors can provide insight into the oxidative chemistry at play. The analysis is extended to particle-phase data to show further classifications of model systems based on both broad and detailed SOA composition measurements.

The methodology described and the results presented (supported by findings obtained from zero-dimensional box modelling), indicate that there is some potential that the approach could ultimately provide the foundations for a framework onto which it would be possible to map the chemistry and oxidation characteristics of ambient air measurements. This could in turn allow “pattern” typing and source origination for certain complex air matrices and provide a snapshot of the reactive chemistry at work, lending insight into the type of chemistry driving the compositional change of the contemporary atmosphere. There are similarities between this approach to discovery science in the atmosphere and metabolomics strategies in biology (e.g. Sousa et al., 2013; Kuppusami et al., 2014).

Deconvolution of complex atmospheric datasets

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



## 2 Experimental

### 2.1 Choice of precursors

Six different BVOCs and one anthropogenic VOC were chosen for analysis. The target compounds, their structures and reaction rate constants with respect to OH and O<sub>3</sub> are given in Table 1. The BVOCs were chosen according to their atmospheric prevalence, structure and contrasting photooxidative reaction pathways; all have previously been shown to form SOA under simulation chamber conditions (e.g. Lee et al., 2006; Alfara et al., 2013) and references therein). Isoprene is a C<sub>5</sub> diene that accounts for around 62 % (~ 594 Tgyr<sup>-1</sup>) of total annual non-methane BVOC emissions (Sindelarova et al., 2014). After isoprene, monoterpenes (C<sub>5</sub>H<sub>16</sub>) have the next largest annual emission rate, they account for around 11 % (~ 95 Tgyr<sup>-1</sup>) of total annual non-methane BVOC emissions (Sindelarova et al., 2014).  $\alpha$ -pinene and limonene were chosen for analysis here alongside isoprene, the former acting as a model system to represent bicyclic monoterpenes, the later to represent monocyclic diene terpenes. In this work,  $\alpha$ -pinene and limonene together generically represent (and are referred to hereafter as) “cyclic” monoterpenes (i.e. monoterpenes that contain one six-member carbon ring). In order to explore the chemistry of non-cyclic monoterpenes, myrcene, an acyclic triene monoterpene, was also included, as was the structurally similar acyclic diene OVOC, linalool. In this work, myrcene and linalool together generically represent (and are referred to hereafter as) “straight chain” monoterpenes/BVOCs (note: linalool is not technically a monoterpene, but does contain the same carbon backbone as myrcene, consequently it is expected to exhibit similar photooxidative chemistry). Finally,  $\beta$ -caryophyllene was included to represent sesquiterpenes, which have annual emissions of the order 20 Tgyr<sup>-1</sup> (Sindelarova et al., 2014). In order to test the ability of the methodology to distinguish between biogenic and anthropogenic systems, toluene was also included. Toluene is often used as a model system to act as a proxy for aromatic species in general (e.g. Bloss et al., 2005). For contrasting plant mesocosm sys-

tems, *Ficus benjamina* and *Ficus cyathistipula* (fig) and *Betula pendula* (birch) species were chosen to represent tropical rainforest and European environs, respectively.

In general, the VOC precursors employed have roughly similar reaction rate constants with respect to OH and O<sub>3</sub>, e.g. limonene, myrcene, linalool and  $\beta$ -caryophyllene all have atmospheric lifetimes with respect to OH of the order 40–50 min (Alfarra et al., 2013; Atkinson and Arey, 2003b).  $\beta$ -caryophyllene has the shortest lifetime with respect to O<sub>3</sub> (ca. 2 min) and isoprene and  $\alpha$ -pinene have the longest lifetimes with respect to both OH and O<sub>3</sub>, e.g. isoprene and  $\alpha$ -pinene have atmospheric lifetimes with respect to OH of the order 1.4–2.7 h (Alfarra et al., 2013; Atkinson and Arey, 2003b). In order to ensure the various systems had progressed sufficiently down their respective photooxidative reaction pathways, the experiment duration was set to be sufficiently long that the majority of the precursor had been consumed by the conclusion of the experiment.

## 2.2 Chamber infrastructure

Experiments were carried out across three different European environmental simulation chamber facilities over a number of separate campaigns. The chambers used, included (1) The University of Manchester Aerosol Chamber (MAC), UK (Alfarra et al., 2012); (2) The European Photoreactor (EUPHORE), ES (Becker, 1996) and (3) The Paul Scherrer Institut Smog Chamber (PSISC), CH (Paulsen et al., 2005). A brief technical description of each facility is given in Table 2.

## 2.3 Experiment design

Table 1 provides a summary of the experiments conducted, which can be divided into three separate categories, (1) photooxidation, indoor chamber (Wyche et al., 2009; Alfarra et al., 2012, 2013), (2) photooxidation, outdoor chamber (Bloss et al., 2005; Camredon et al., 2010) and (3) mesocosm photooxidation, indoor chamber (Wyche et al., 2014). In each case the reaction chamber matrix comprised a temperature ( $T = 292$ – $299$  K) and humidity (49–84 % for photooxidation, indoor chamber and < 2–6 % for

## Deconvolution of complex atmospheric datasets

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



photooxidation, outdoor chamber) controlled synthetic air mixture. For all experiments the chamber air matrix also contained a pre-defined initial quantity of NO and NO<sub>2</sub> (VOC/NO<sub>x</sub> ratios in the range 0.6–20, but typical ~ 2). The VOC precursor was introduced into the reaction chamber in liquid form via a heated inlet. In the case of the mesocosm photooxidation experiments, a known volume of air containing the precursor VOCs was transferred to the reaction chamber from a separate, illuminated plant chamber, which contained several tree specimens. For the indoor chamber systems, the experiments were initiated, after introduction of all reactants, by the switching on of artificial lights. For the outdoor chamber systems, the opening of the chamber cupola marked the start of the experiment. Experiments were typically run for 4–6 h.

## 2.4 Instrumentation

CIR-TOF-MS was used to make *real-time* (i.e. 1 min) measurements of the complex distribution of volatile organic compounds ( $\Sigma$ VOC, i.e. the sum of VOCs, oxygenated VOCs – OVOCs and nitrated VOCs – NVOCs) produced in the gas-phase during oxidation of each parent compound. In brief, the CIR-TOF-MS comprises a temperature controlled ( $T = 40^\circ\text{C}$ ) ion source/drift cell assembly coupled to an orthogonal time-of-flight mass spectrometer equipped with a reflectron array (Kore Technology, UK). Proton Transfer Reaction (PTR) from hydronium ( $\text{H}_3\text{O}^+$ ) and hydrated hydronium ( $\text{H}_3\text{O}^+ \cdot (\text{H}_2\text{O})_n$ ) was employed as the ionisation technique during all experiments (Jenkin et al., 2012). Further details regarding the CIR-TOF-MS can be found in Blake et al. (2004) and Wyche et al. (2007).

Aerosol samples were collected on 47 mm quartz fibre filters at the end of certain experiments and the water-soluble organic content was extracted for analysis using LC-MS/MS. Reversed phase LC separation was achieved using an HP 1100 LC system equipped with an Eclipse ODS-C18 column with 5  $\mu\text{m}$  particle size (Agilent, 4.6 mm  $\times$  150 mm). Mass spectrometric analysis was performed in negative ionisation mode using an HCT-Plus ion trap mass spectrometer with electrospray ionisation (Bruker Daltonics GmbH). Further details can be found in Hamilton et al. (2003).

## Deconvolution of complex atmospheric datasets

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



## Deconvolution of complex atmospheric datasets

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



For several experiments, *real-time* broad chemical characterisation of the SOA was made using a cTOF-AMS (Aerodyne Research Inc., USA). The cTOF-AMS was operated in standard configuration, taking both mass spectrum (MS) and particle time-of-flight (PTOF) data; it was calibrated for ionisation efficiency using 350 nm monodisperse ammonium nitrate particles, the vapouriser was set to  $\sim 600^\circ\text{C}$  and a collection efficiency value of unity was applied (Alfarra et al., 2006). For further details, refer to Drewnick et al. (2005) and Canagaratna et al. (2007).

Each chamber was additionally instrumented with on-line chemiluminescence (pholytic  $\text{NO}_2$ )  $\text{NO}_x$  analysers, UV photometric  $\text{O}_3$  detectors, and scanning mobility particle sizers and condensation particle counters for aerosol size and number concentration, as well as temperature, pressure and humidity monitors. For full details regarding the various instrument suites employed at each chamber see Alfarra et al. (2012), Paulsen et al. (2005), Camredon et al. (2010) and references therein.

Filter and cTOF-AMS data were collected only during photooxidation experiments conducted at the MAC. Repeat experiments conducted at the MAC were carried out under similar starting conditions (e.g. VOC/ $\text{NO}_x$  ratio Alfarra et al., 2013).

### 2.5 Model construction

In order to aid analysis, the composition and evolution of the gas-phase components of the  $\alpha$ -pinene chamber system were simulated using a chamber optimised photochemical box model incorporating the comprehensive  $\alpha$ -pinene atmospheric oxidation scheme extracted from the Master Chemical Mechanism website (Jenkin et al., 1997, 2012; Saunders et al., 2003; <http://mcm.leeds.ac.uk/MCM>). The  $\alpha$ -pinene mechanism employed (along with an appropriate inorganic reaction scheme) contained approximately 313 species and 942 different reactions. The box model employed also incorporated a series of “chamber specific” auxiliary reactions adapted from Bloss et al. (2005), Zador et al. (2006) and Metzger et al. (2008) in order to take into account background chamber reactivity. Photolysis rates were parameterised for the PSI chamber and constrained using measured values of ( $j(\text{NO}_2)$ ). All simulations were

run at 295 K and 50% relative humidity. NO, NO<sub>2</sub>, HONO and  $\alpha$ -pinene were either initialised or constrained, depending on the scenario investigated. For further details see Rickard et al. (2010).

### 3 Data analysis

#### 3.1 Data processing

All CIR-TOF-MS data were recorded at a time resolution of 1 min. In order to remove the time dimension and simultaneously increase detection limit, the individual mass spectra were integrated over the entire experiment; as such no account is taken of overall reaction time in the CIR-TOF-MS analysis. Removing the time dimension acts to reduce the dimensionality of the data, whilst maintaining the central characteristic spectral fingerprints produced by the photooxidation process. On average across all experiments studied, 98% of the precursor had been consumed by the conclusion of the experiment; hence it is assumed that sufficient reaction took place in each instance to provide summed-normalised mass spectra that fully capture first- and higher-generation product formation.

The resultant summed spectra were normalised to 10<sup>6</sup> primary reagent ion counts (i.e.  $\Sigma(\text{H}_3\text{O}^+ + \text{H}_3\text{O}^+ \cdot (\text{H}_2\text{O})_n)$ ). Similarly normalised background spectra (recorded prior to injection of the precursor) were then subtracted from the summed-and-normalised experiment spectra. The  $65 < m/z < 255$  channels of the background removed spectra were extracted to comprise the region of interest. These ions tend to carry the most analyte-specific information, with lower  $m/z$  features tending to comprise either generic fragment ions that provide little chemical information (Blake et al., 2006) and/or small compounds emitted from illuminated chamber walls (e.g. Bloss et al., 2005; Zador et al., 2006; Metzger et al., 2008). These extracted data were refined further by the application of a Mann–Whitney test (see Statistical Analysis for details), leaving residual spectra that comprised only the integrated-over-time signals corresponding to the VOC

## Deconvolution of complex atmospheric datasets

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



precursor and any reactive intermediate and product VOCs formed within the chamber during the experiment. Finally, the signal counts (in units of normalised counts per second; ncps) in each mass channel of the residuals, were expressed as a percentage of the total ion count in the refined region of interest.

The LC-MS/MS signal intensity data for the region  $51 < m/z < 599$  were extracted for analysis. For the AMS data, a 10 min average was produced at 4 h after lights on (roughly around the time when SOA mass had reached a peak and towards to the end of the experiment) and the region  $40 < m/z < 150$  (again the region carrying the most information; Alfara et al., 2006) was extracted. Similar to the gas-phase data sets, the LC-MS/MS and AMS data were filtered using a Mann–Whitney test. Finally, for each data set all signal counts were expressed as a percentage of the total ion count in the respective  $m/z$  region of interest.

### 3.2 Statistical analysis

Before any multivariate analysis was conducted, the processed CIR-TOF-MS, LC-MS/MS and AMS spectra were first filtered to remove unwanted data that were deemed to not be statistically significant. In order to do this, the mass spectra were initially grouped by structure of the precursor employed, giving seven separate groups for the CIR-TOF-MS data and three groups (owing to the smaller number of precursor species investigated) for the LC-MS/MS and AMS data, respectively. A two-sided Mann–Whitney test was then used to assess whether signals reported in individual mass channels were significantly different from the corresponding signals measured during blank experiments. SPSS V20 (IBM, USA) was used for the analysis. A  $p$  value of  $< 0.05$  was considered statistically significant. The final summed-normalised and filtered spectra were then subjected to a series of multivariate statistical analysis techniques in order to probe the underlying chemical information. PLS-Toolbox (Eigenvector Research Inc., USA) operated in MatLab (Mathworks, USA; PLS-Tool Box) was used for the analysis.

## Deconvolution of complex atmospheric datasets

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



**Deconvolution of complex atmospheric datasets**

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



To begin with, to reduce the data and identify similarities between the precursor oxidation systems, a PCA was conducted on the BVOC dataset and the model generated was then employed to map the reactivity of fig and birch tree mesocosm systems and to investigate the fit of a typical anthropogenic system (toluene) into the PCA space (both introduced into the model as test datasets). An *unsupervised pattern recognition*, hierarchical cluster analysis was also conducted on the data and a dendrogram produced to test relatedness, support the PCA and help interpret the precursor class separations achieved. The dendrogram was constructed using PCA scores, the centroid method and Mahalanobis distance coefficients. Finally, a *supervised pattern recognition* PLS-DA analysis was employed as a check for false-positives and as a quantitative classification tool to test the effectiveness of classification of the various systems in the model.

For the superposition of “classification” confidence limits onto the results of the PCA and HCA and for classification discrimination in the PLS-DA, prior to analysis the experiments were grouped according to the structure of the precursor investigated. Group 1 = isoprene (hemiterpene) and group 2 =  $\alpha$ -pinene and limonene (both cyclic monoterpenes with an endocyclic double bond). Although limonene also has an exocyclic double bond in a side chain, we justify this classification on account of the endocyclic double bond in limonene being much more reactive towards ozone and slightly more reactive towards OH (Calvert et al., 2000). Group 3 =  $\beta$ -caryophyllene (sesquiterpene) and group 4 = myrcene (straight chain monoterpene) and linalool (straight chain OVOC). Strictly speaking, linalool is an OVOC (structure  $C_{10}H_{18}O$ ) and not a monoterpene (structure  $C_{10}H_{16}$ ), however we justify this grouping on account of both myrcene and linalool comprising primary BVOCs (often co-emitted; Bouvier-Brown et al., 2009; Kim et al., 2010; Wyche et al., 2014) with certain structural similarities.



## Deconvolution of complex atmospheric datasets

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



The  $m/z$  loadings of the PCA allow us to understand how the spectral fingerprints of the different terpene oxidation systems are grouped/separated by the PCA model. The first set of ions that contribute to separation of the different terpene systems comprises the protonated parent ions ( $MH^+$ ) of the precursors themselves (and major fragments thereof), i.e.  $m/z$  69 for isoprene, 137 (and fragment 81) for all monoterpenes (regardless of structure) and 205 for  $\beta$ -caryophyllene. Important contributions are to be expected from the respective parent-ions (being the basis for the use of chemical-ionisation mass spectrometry as an analyser of gas mixtures, Blake et al., 2009). Our purpose here goes beyond identification of precursor and intermediate VOCs to an interpretation of reaction pathways in complex mixtures. In doing this, a certain amount of disambiguation of isobaric compounds becomes possible; indeed, as discussed in more detail below, Fig. 2 clearly shows separation between cyclic and non-cyclic monoterpene groups, both of which have precursors of molecular weight (MW)  $136\text{ g mol}^{-1}$ .

Moving past the precursors into the detailed chemical information provided by the oxidation products formed within the chamber, we can see from Fig. 2 that amongst others,  $m/z$  71 (methyl vinyl ketone and methacrolein), 75 (hydroxy acetone), 83 (e.g. 3-methyl furan) and 87 ( $C_4$ -hydroxycarbonyls/methacrylic acid) all contribute to separation of the isoprene group, and  $m/z$  237 ( $\beta$ -caryophyllon aldehyde) and 235 and 253 ( $\beta$ -caryophyllene secondary ozonide and isomers thereof) to that of the  $\beta$ -caryophyllene group. The monoterpene groupings are influenced by the presence of  $m/z$  107, 151 and 169 (primary aldehydes, pionaldehyde and limonondehyde) and 139 (primary ketone, limonaketone) ions in their mass spectra. Helping to separate the straight chain from cyclic monoterpenes are  $m/z$  95 and 93, dominant features in both the myrcene and linalool spectra (relative abundance 10–24 % for  $m/z$  93).  $m/z$  93 has previously been identified as a major fragment ion of first generation myrcene and linalool products 4-vinyl-4-pentenal and 4-hydroxy-4-methyl-5-hexen-1-al, respectively (Shu et al., 1997; Lee et al., 2006). Note, for clarity within Fig. 2, the scale has been set to show the bulk of the data, hence precursor parent ions and  $m/z$  71 are not shown.







**Deconvolution of  
complex atmospheric  
datasets**

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



MS spectra, with distinct clusters of cyclic monoterpenes, straight chain monoterpenes and sesquiterpenes. From inspection of the loadings components of the biplot (Fig. 7a), we can see that  $m/z$  237 (3-[2,2-dimethyl-4-(1-methylene-4-oxo-butyl)-cyclobutyl]-propanoic acid), 251 ( $\beta$ -caryophyllonic acid), 255 (4-(2-(2-carboxyethyl)-3,3-dimethylcyclobutyl)-4-oxobutanoic acid), 267 ( $\beta$ -14-hydroxycaryophyllonic acid and  $\beta$ -10-hydroxycaryophyllonic acid) and 271 (4-(2-(3-hydroperoxy-3-oxopropyl)-3,3-dimethylcyclobutyl)-4-oxobutanoic acid or 4-(2-(2-carboxy-1-hydroxyethyl)-3,3-dimethylcyclobutyl)-4-oxobutanoic acid), are amongst those ions dominant in classifying the  $\beta$ -caryophyllene. For further details regarding  $\beta$ -caryophyllene oxidation products, see Hamilton et al. (2011) and Jenkin et al. (2012) and Sect. 5. Of this set of oxidation products,  $\beta$ -caryophyllonic acid is common between the gas- (i.e.  $m/z$  253) and particle- (i.e.  $m/z$  251) phases.

Similarly, those ions (compounds) significant in isolating the cyclic monoterpenes include,  $m/z$  169 (pinalic-3-acid, ketolimononaldehyde and limonic acid), 183 (pinonic acid, limonic acid and 7-hydroxylimononaldehyde) and 185 (pinic acid, limonic acid), of which only those compounds of  $m/z$  169 were observed to be of significant contribution to the gas-phase composition (observed as  $m/z$  171; relative contribution as high as 1–5% during  $\alpha$ -pinene experiments). For further details regarding  $\alpha$ -pinene and limonene oxidation products, see for example Jenkin (2004), Lee et al. (2006), Camredon et al. (2010) and Hamilton et al. (2011). Comparatively little information is available on the speciated composition of myrcene and linalool SOA, however, from Fig. 7a it is clear that somewhat larger mass compounds are important in classifying straight chain monoterpenes, e.g.  $m/z$  321 (adduct ion  $[M-H_2 + FA + Na]^-$   $M = 254$  Da; potential formulae –  $C_{12}H_{14}O_6$ , six double bond equivalents or  $C_{13}H_{18}O_5$ , five double bond equivalents; indicative of oligomer formation), 325, 322 (the C13 peak for the  $m/z$  321 ion), 227 ( $C_{10}H_{11}O_6$ ), 215 ( $C_{10}H_{15}O_5$ ) and 199 ( $C_9H_{11}O_5$ ). Compounds of such high molecular weight were not observed in the concomitant gas-phase spectra.

As with the PCA, the dendrogram produced via cluster analysis of the LC-MS/MS particle-phase data gave three distinct clusters (Fig. 7b), i.e. cyclic monoterpene,

straight chain monoterpene and sesquiterpene. The corresponding PLS-DA analysis reported 100 % sensitivity in each case and 100 % specificity for all systems except sesquiterpenes (i.e.  $\beta$ -caryophyllene = 83 %), suggesting a good level of model classification for the three types of terpene systems studied.

Despite utilising the somewhat destructive electron impact (EI) ionisation technique, the cTOF-AMS produces spectra of sufficient chemical detail such that the PCA and HCA are able to successfully differentiate between the groups of terpenes tested (Fig. 8a and b). However, unlike the outputs from the CIR-TOF-MS and LC-MS/MS PCA's, the cyclic and straight chain monoterpenes in the AMS PCA do not group into two distinct classes, instead they tend to group in their species-specific sub-classes within the upper half of the PCA space. Indeed, the PLS-DA gave 100 % sensitivity and specificity for the cyclic monoterpenes and sesquiterpenes, but only 75 % sensitivity for the straight chain monoterpenes, suggesting that the model does less well at assigning myrcene and linalool cTOF-AMS spectra to their defined class.

As can be seen from inspection of Fig. 8a,  $\alpha$ -pinene, limonene and linalool tend in general to cluster towards the upper and right regions of the PCA space, primarily owing to the significant presence of  $m/z$  43 and to a lesser extent  $m/z$  44, in their spectra; both ions constituting common fragments observed in AMS of SOA (Alfarra et al., 2006). During such chamber experiments, the  $m/z$  43 peak tends to comprise the  $\text{CH}_3\text{CO}^+$  ion, originating from oxidised compounds containing carbonyl functionalities; it is usually representative of freshly oxidised material and semi-volatile oxygenated organic aerosol (SV-OOA; Alfarra et al., 2006).

From further inspection of the loadings bi-plot (Fig. 8a) we see that the four sesquiterpene ( $\beta$ -caryophyllene) experiments cluster towards the lower left hand quadrant, their clustering heavily influenced by the presence of  $m/z$  41 in their spectra as well as  $m/z$  55, 79 and 95. In EI-AMS,  $m/z$  41 comprises the unsaturated  $\text{C}_3\text{H}_5^+$  fragment (Alfarra et al., 2006). As well as being influenced by the  $m/z$  41 ion, the myrcene cluster (situated in the region of both the  $\alpha$ -pinene and  $\beta$ -caryophyllene clusters in the PCA space) is also influenced by  $m/z$  44, i.e. most likely the  $\text{CO}_2^+$  ion. In this instance  $m/z$

## Deconvolution of complex atmospheric datasets

K. P. Wyche et al.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)



100's ppbV; Paulot et al., 2009b; Surratt et al., 2010, 2006), and ISOPOOH and IEPOX ( $\text{MH}^+ - \text{H}_2\text{O} = m/z$  101) under "low  $\text{NO}_x$ " conditions. For the "high  $\text{NO}_x$ " isoprene experiments conducted here, besides  $m/z$  71, i.e. MACR (measured together with methyl vinyl ketone),  $m/z$  87, 85, 83 and 75 i.e. (tentatively assigned to be)  $\text{C}_4$ -hydroxycarbonyls/methacrylic acid, ISOPN,  $\text{C}_5$ -hydroxy carbonyls ( $\text{C}_5\text{HC}$  in Fig. 9)/3-methyl furan (3-MF) and hydroxy acetone, respectively, were significant in classifying the isoprene group; MPAN at the  $m/z$  103 ion was only a minor contributor. It should be noted that in theory, both HMML and MAE ( $\text{MH}^+ = m/z$  103) may produce fragment ions of  $m/z$  85 (i.e.  $\text{MH}^+ - \text{H}_2\text{O}$ ) following PTR ionisation, however without further detailed characterisation we are unable at this stage to postulate their fractional contribution to the measured  $m/z$  85 signal.

Depending on the chemistry involved (Fig. 9), potential SOA forming monoterpene products will either be (six-member-) ring retaining (e.g. from reaction with OH) or (six-member-) ring cleaved (e.g. from reaction with OH or  $\text{O}_3$ ), producing gas-phase spectra with mid MW  $\text{C}_9$  and  $\text{C}_{10}$  oxygenated (and nitrated in the presence of  $\text{NO}_x$ ) products (e.g. Kamens and Jaoui, 2001; Larsen et al., 2001; Capouet et al., 2004; Yu et al., 2008; Camredon et al., 2010; Eddingsaas et al., 2012b). Both (six-member-) ring retaining and (six-member-) ring-opening products have been observed in monoterpene SOA (e.g. Yu et al., 1999; Larsen et al., 2001; Camredon et al., 2010), with the latter generally being dominant in terms of abundance (Camredon et al., 2010). Furthermore, (six-member-) ring-opening products are believed to undergo chemistry within the aerosol to form relatively low O:C ratio oligomers (e.g. 10-hydroxy-pinonic acid-pinonic acid dimer, O:C = 7 : 19) (Gao et al., 2004; Tolocka et al., 2004; Camredon et al., 2010).

OH will react with straight chain monoterpenes, such as myrcene, primarily by addition to either the isolated or the conjugated double bond system. Reaction at the isolated C=C bond can proceed via fragmentation of the carbon backbone, producing acetone and mid MW, unsaturated  $\text{C}_7$  OVOCs (and/or NVOCs, depending on  $\text{NO}_x$  levels). Reaction at the conjugated double bond system in myrcene would be expected to form

## Deconvolution of complex atmospheric datasets

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



formaldehyde in conjunction with either a C<sub>9</sub> aldehyde or C<sub>9</sub> ketone. Structure activity relationships (SARs) predict that the conjugated double bond system accounts for almost half of the OH reactivity. The conjugated double bond would therefore be expected to have a partial rate coefficient of the order  $1 \times 10^{-10}$  (i.e. similar to OH + isoprene) (Atkinson and Arey, 2003b). Consistent with this, the reported yields of acetone and formaldehyde from OH + myrcene are similar (Atkinson and Arey, 2003b), suggesting that the isolated double bond and the conjugated double bond system have comparable OH reactivity, as such we would expect C<sub>9</sub> and C<sub>7</sub> co-products to be formed in comparable yields. However, with a significant fraction of reactions with OH leading to the loss of three carbon atoms from the parent structure, the straight chain monoterpene gas-phase spectra tend to contain fewer features of MW greater than that of the precursor and more mid MW features. It tends to be these mid MW features, such as *m/z* 111 and 93 (e.g. 4-vinyl-4-pentenal, MYR 1.2 in Fig. 9, MH<sup>+</sup> and MH<sup>+</sup>-H<sub>2</sub>O, respectively) and 113 and 95 (e.g. 2-methylenepentanedial MH<sup>+</sup> and MH<sup>+</sup>-H<sub>2</sub>O, respectively) that assist in the classification of the straight chain monoterpene experiments within the statistical space. Besides these ions, *m/z* 139 (primary myrcene C<sub>9</sub> aldehyde and/or C<sub>9</sub> ketone product) also assists in separating the myrcene spectra from those of  $\alpha$ -pinene.

By comparing both the gas- and particle-phase cyclic monoterpenes in Figs. 2 and 7a, it is evident that the dominant loadings represent compounds of similar MW, i.e. 169, 151 and 107 (primary aldehyde product, e.g. pinonaldehyde- PINAL in Fig. 9, parent ion and fragments thereof) and 139 (primary ketone product parent ion) for the gas-phase and 187, 185, 183 and 169 for the particle-phase. Conversely, for the straight chain monoterpene experiments the major gas-phase loadings represent compounds of significantly smaller MW than their particle-phase counterparts, i.e. 113 and 95 and 111 and 93, compared to 325, 322, 321, 227 and 215. Indeed, the straight chain monoterpene LC-MS/MS spectra contained on average  $\sim 10\%$  more signal  $> 250$  Da than the cyclic monoterpene spectra. Also, the composition of the ions observed in the straight chain monoterpene LC-MS/MS spectra suggests that the SOA particles contained both oligomers and highly oxidized species, with the C<sub>10</sub> backbone intact (i.e.

## Deconvolution of complex atmospheric datasets

K. P. Wyche et al.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

**Deconvolution of  
complex atmospheric  
datasets**

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



O : C = 0.6), similar in structure to (but a little less oxidised than) extremely low volatility organic vapours (ELV-VOC), which have been observed previously in significant yield from  $\alpha$ -pinene and limonene (as well as 6-nonenal) ozonolysis chamber experiments in the absence of an OH scavenger, as well as boreal forests in Finland (Ehn et al., 2014). Further evidence to elucidate the type of SOA formed from the oxidation of straight chain monoterpenes can be obtained from investigation of the grouping of myrcene spectra in the cTOF-AMS PCA (Fig. 8a). In the hour-4 cTOF-AMS PCA loadings bi-plot, we see that the grouping of the myrcene spectra is influenced somewhat by both  $m/z$  41 and 44, indicating the presence of LV-OOA in the SOA, potentially a result of oligomerisation or further oxidative heterogeneous chemistry involving reaction at remaining C=C double bond sites.

$\beta$ -caryophyllene readily forms particulate matter on oxidation (Alfarra et al., 2012), with reaction predominantly at one of the two C=C sites (e.g. with OH or O<sub>3</sub>, although O<sub>3</sub> attack occurs almost exclusively at the endocyclic double bond, Jenkin et al., 2012), yielding relatively low vapour pressure, unsaturated and oxygenated primary products (Fig. 9), which have significant affinity for the particle-phase (Jenkin et al., 2012). A further oxidation step involving the second C=C site can result in increased oxygen (and/or nitrogen, depending on NO<sub>x</sub> conditions) content, yet with little, if any reduction in the original C number. As with the cyclic monoterpene PCAs, the CIR-TOF-MS and LC-MS/MS PCA bi-plots demonstrate similarities in terms of classifying  $\beta$ -caryophyllene oxidation and SOA formation with comparable MW species, e.g. primary products  $\beta$ -caryophyllon aldehyde (MW 236, BCAL in Fig. 9) and  $\beta$ -caryophyllene secondary ozonide in the gas-phase (MW 252, BCSOZ in Fig. 9),  $\beta$ -caryophyllonic acid (MW 252, C141CO<sub>2</sub>H in Fig. 9) in both phases and secondary product  $\beta$ -nocaryophyllinic acid (MW 254, C131CO<sub>2</sub>H in Fig. 9) in the particle-phase. In the hour-4 cTOF-AMS PCA scores plot, the myrcene and  $\beta$ -caryophyllene clusters are located adjacent to one another, with  $\beta$ -caryophyllene classification also influenced by the  $m/z$  41 peak, which similar to myrcene SOA for example, is indicative of higher oxidized content (Alfarra et al., 2012), a result of either the partitioning of higher gen-

eration gas-phase products or heterogeneous oxidation of condensed first or second generation products.

## 6 Atmospheric relevance and future directions

Having successfully used the mechanistic fingerprints in the chamber data to construct descriptive statistical models of the gas- and particle-phases, and having applied the methodology to map mesocosm environments, a next logical step would be to use this detailed chemical knowledge to investigate ambient VOC and SOA composition data in an attempt to help elucidate and deconvolve the important chemistry controlling the gas- and particle-phase composition of inherently more complex real world environments.

If ambient biogenic gas/particle composition spectra of unknown origin, uncertain speciated composition and/or a high level of detail and complexity were to be mapped onto the relevant statistical model (i.e. introduced as a separate test set), their resultant vector description in the statistical space would provide information regarding the type of precursors present and the underlying chemical mechanisms at play, as exemplified by the classifying of the mesocosm experiments by the fraction of isoprene, monoterpene and sesquiterpene chemistry in the experimental fingerprints. Furthermore, as shown by the mapping of toluene photooxidation experiments into a separate and distinct cluster, the methodology is potentially able to be robust with respect to other chemical compositions expected for a “real world” environment that is significantly impacted by both anthropogenic and biogenic emissions (e.g. Houston, USA and the Black Forest – Munich, DE). This capability is important when attempting to understand the complex interactions that exist between urban and rural atmospheres and when attempting to understand VOC and SOA source identification.

One potential problem in moving from simulation chamber data to “real world” systems, would be the applicability of using “static” experimental spectra (i.e. time aver-

### Deconvolution of complex atmospheric datasets

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion





**Deconvolution of  
complex atmospheric  
datasets**

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



Figure 10c and d shows the results from scenario (2). Figure 10c clearly shows the second  $\alpha$ -pinene injection on top of the evolving matrix and the resultant system evolution. Figure 10d shows the “difference model mass spectra” between scenarios (1) and (2), from which it can clearly be seen that there is very little difference between the spectra of the basic model and the “spiked” system. The difference in “mass channel” relative abundance ( $\Delta MC$ ) is generally  $\leq 2\%$ , with the exceptions of MWs 168 and 186. MW 168 primarily comprises pinonaldehyde, with a  $\Delta MC$  of around  $-6\%$ ; pinonaldehyde is a primary product and is slightly lower in relative abundance in scenario (2) owing to the longer reaction time employed and the greater proportion of pinonaldehyde reacted. MW 186 comprises a number of primary and secondary products and has a  $\Delta MC$  of roughly  $+3\%$ .

The results from model scenario (3) are given in Fig. 10e and f. As with scenario (2), there is no dramatic difference between the simulated mass spectra of scenario (3) and the base-case scenario (1). In this instance  $\Delta MC$  is generally  $\leq \pm 5\%$ , with the exceptions of MWs 136 and 168 and MWs 121 and 245. The relative abundance of the precursor is lower in this case on account of the constraining method employed and once again the relative abundance of pinonaldehyde is slightly lower due to the longer reaction time. MW 121 solely comprises PAN and MW 245 primarily comprises a  $C_{10}$  tertiary nitrate ( $C_{10}H_{15}NO_6$ , MCM designation: C106NO3). Both species are slightly elevated with respect to the base-case in scenario (3) owing to the longer reaction time and the continual input of OH and NO into the model in the form HONO.

Scenarios (2) and (3) represent complex mixtures with overlapping reaction coordinates, each one step closer to a “real world” case than scenario (1) and the chamber data employed within this work. However, despite the increase in complexity of the scenarios, both exhibit very little compositional difference to the base-case scenario and hence the chamber data employed in this work. These results give some confidence that despite being constructed from summed simulation chamber data, the statistical models employed here represents a solid framework onto which real atmosphere spectra could be mapped and interpreted.

## Deconvolution of complex atmospheric datasets

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



A further step in increasing complexity and hence a further step towards the “real world” system, would be the addition of other (potentially unidentified) precursors to the simulation, which may be at different stages of oxidation or have passed through different reactive environments. Further increases in complexity, beyond the analysis discussed here, will form the focus of future work.

## 7 Conclusions

A chemometric dimension reduction methodology, comprising PCA, HCA and PLS-DA has been successfully applied for the first time to complex gas- and particle-phase composition spectra of a wide range of BVOC and mesocosm environmental simulation chamber photooxidation experiments. The results show that the oxidized gas-phase atmosphere (i.e. the integrated reaction coordinate) of each different structural type of BVOC can be classified into a distinct group according to the controlling chemistry and the products formed. Indeed, a major strength of the data analysis methodology described here, lies in the decoding of mechanisms into pathways and consequently linking the pathways to precursor compounds. Furthermore, the methodology was similarly able to differentiate between the types of SOA particles formed by each different class of terpene, both in the detailed and broad chemical composition spectra. In concert, these results show the different SOA formation chemistry, starting in the gas-phase, proceeding to govern the differences between the various terpene particle compositions.

The ability of the methodology employed here to efficiently and effectively “data mine” large and complex datasets becomes particularly pertinent when considering that modern instrumentation/techniques produce large quantities of high-resolution temporal and speciated data over potentially long observation periods. Such statistical mapping of organic reactivity offers the ability to simplify complex chemical datasets and provide rapid and meaningful insight into detailed reaction systems comprising hundreds of reactive species. Moreover, the demonstrated methodology has the potential to as-

sist in the evaluation of (chamber and real world) modelling results, providing easy to use, comprehensive observational metrics with which to test and evaluate model mechanisms and outputs and thus help advance our understanding of complex organic oxidation chemistry and SOA formation.

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## Deconvolution of complex atmospheric datasets

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



**Deconvolution of complex atmospheric datasets**

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



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**Deconvolution of  
complex atmospheric  
datasets**

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



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- 30

**Deconvolution of  
complex atmospheric  
datasets**

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



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**Deconvolution of  
complex atmospheric  
datasets**

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



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**Deconvolution of  
complex atmospheric  
datasets**

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



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**Deconvolution of  
complex atmospheric  
datasets**

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



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**Deconvolution of  
complex atmospheric  
datasets**

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



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**Deconvolution of  
complex atmospheric  
datasets**

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



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**Deconvolution of  
complex atmospheric  
datasets**

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



Crouse, J., Wennberg, P., Wolfe, G. M., Harrold, S., Thornton, J. A., Farmer, D. K., Docherty, K. S., Cubison, M. J., Jimenez, J. L., Frossard, A. A., Russell, L. M., Kristensen, K., Glasius, M., Mao, J. Q., Ren, X. R., Brune, W., Browne, E. C., Pusede, S. E., Cohen, R. C., Seinfeld, J. H., and Goldstein, A. H.: Observational insights into aerosol formation from iso-

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## Deconvolution of complex atmospheric datasets

K. P. Wyche et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

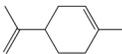
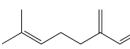
Full Screen / Esc

Printer-friendly Version

Interactive Discussion



**Table 1.** Summary of experiments conducted.

Experiment ID	Precursor	Structure	$k(\text{OH})/k(\text{O}_3)^4/\text{cm}^3 \text{mol}^{-1} \text{s}^{-1}$	Experiment Type (no.)	VOC/ $\text{NO}_x$ Range	RH/% Range
ISOP1–10	isoprene		$9.9 \times 10^{-11}/1.2 \times 10^{-17}$	Photooxidation (10)	1.3–20.0	49–72
APIN1–4 <sup>2,3</sup>	$\alpha$ -pinene		$5.3 \times 10^{-11}/8.4 \times 10^{-17}$	Photooxidation (4)	1.3–2.0 <sup>1</sup>	49–73
LIM1–6 <sup>2,3</sup>	limonene		$1.7 \times 10^{-10}/2.1 \times 10^{-16}$	Photooxidation (6)	1.4–2.0 <sup>1</sup>	50 <sup>1</sup> –82
BCARY1–10 <sup>2,3</sup>	$\beta$ -caryophyllene		$2.0 \times 10^{-10}/1.2 \times 10^{-14}$	Photooxidation (10)	0.6–2.0 <sup>1</sup>	50 <sup>1</sup> –72
MYRC1–2 <sup>2,3</sup>	myrcene		$2.1 \times 10^{-10}/4.7 \times 10^{-16}$	Photooxidation (2)	1.4–1.9	52–54
LINA1–2 <sup>2,3</sup>	linalool		$1.6 \times 10^{-10}/4.5 \times 10^{-16}$	Photooxidation (2)	1.4–2.6	42–47
BIR1–2	birch trees	Multiple emissions <sup>5</sup>	Multiple emissions	Mesocosm Photooxidation (2)	5.5–5.6	73–84
FIG1–2	fig trees	Multiple emissions <sup>5</sup>	Multiple emissions	Mesocosm Photooxidation (2)	2.7–9.4	65–75
TOL1–5	toluene		$3.7 \times 10^{-12}/-$	Photooxidation (5)	1.3–11.6	2–6

<sup>1</sup> Estimated using known volume of reactants injected.

<sup>2</sup> LC-MS/MS filter data available for at least one of these experiments (MAC).

<sup>3</sup> c-TOF-AMS data available for at least one of these experiments (MAC).

<sup>4</sup> From (Atkinson and Arey, 2003b; Sun et al., 2012; Khamaganov and Hites, 2001) and references therein.

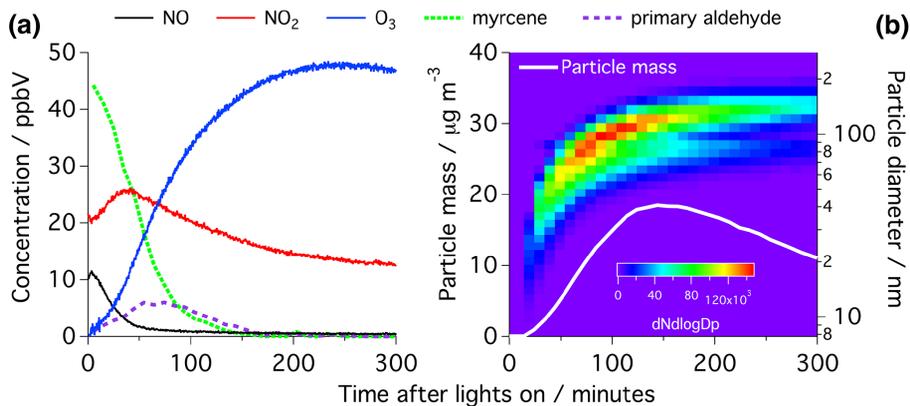
<sup>5</sup> See Wyche et al., 2014.





## Deconvolution of complex atmospheric datasets

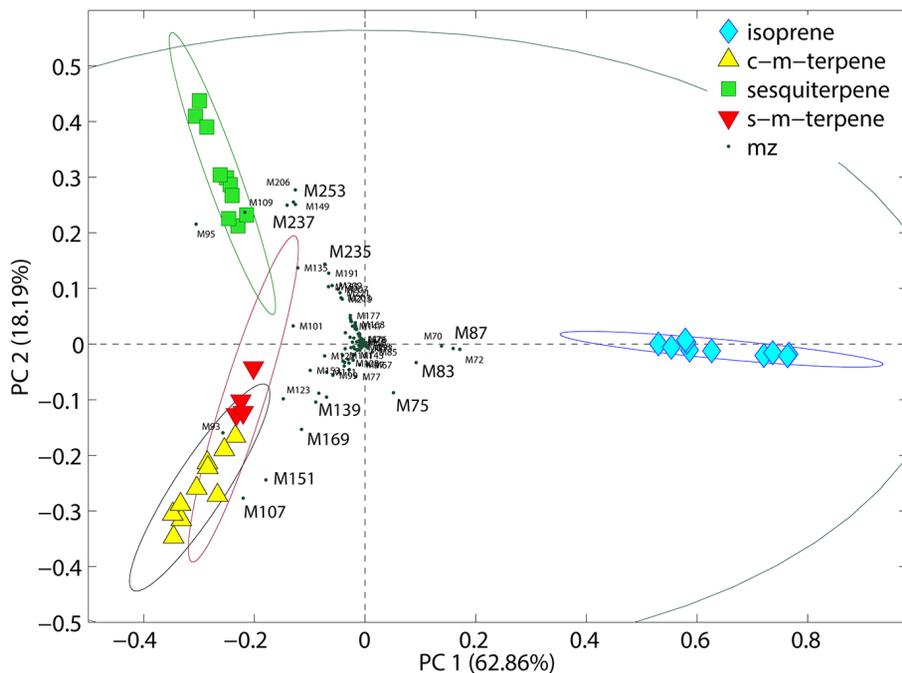
K. P. Wyche et al.



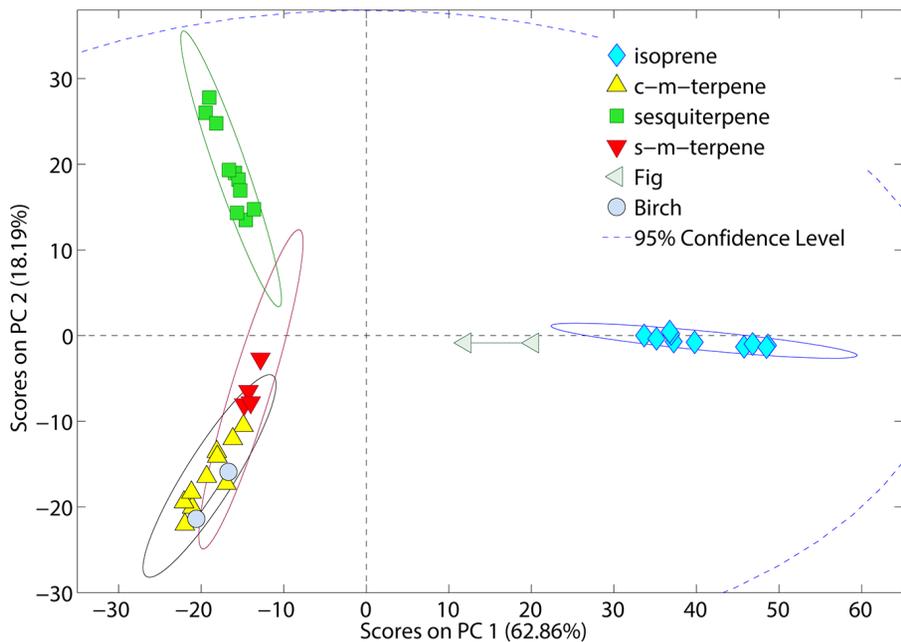
**Figure 1.** (a) NO<sub>x</sub>, O<sub>3</sub>, myrcene and 4-vinyl-4-pentenal (primary aldehyde product) and (b) particle mass (not wall loss corrected and assuming  $\rho = 1.3$ ) and size evolution within the MAC during a typical photooxidation experiment.

## Deconvolution of complex atmospheric datasets

K. P. Wyche et al.



**Figure 2.** PCA loadings bi-plot of the second vs. first principal components derived from the PCA analysis of the isoprene, cyclic monoterpene (“c-m-terpene” in the legend;  $\alpha$ -pinene and limonene), sesquiterpene ( $\beta$ -caryophyllene) and straight chain biogenic (“s-m-terpene” in the legend; myrcene and linalool) chamber data. Classification confidence limits = 95 %. For clarity, the scale has been set to show the bulk of the data, hence precursor parent ions and  $m/z$  71 are not shown.



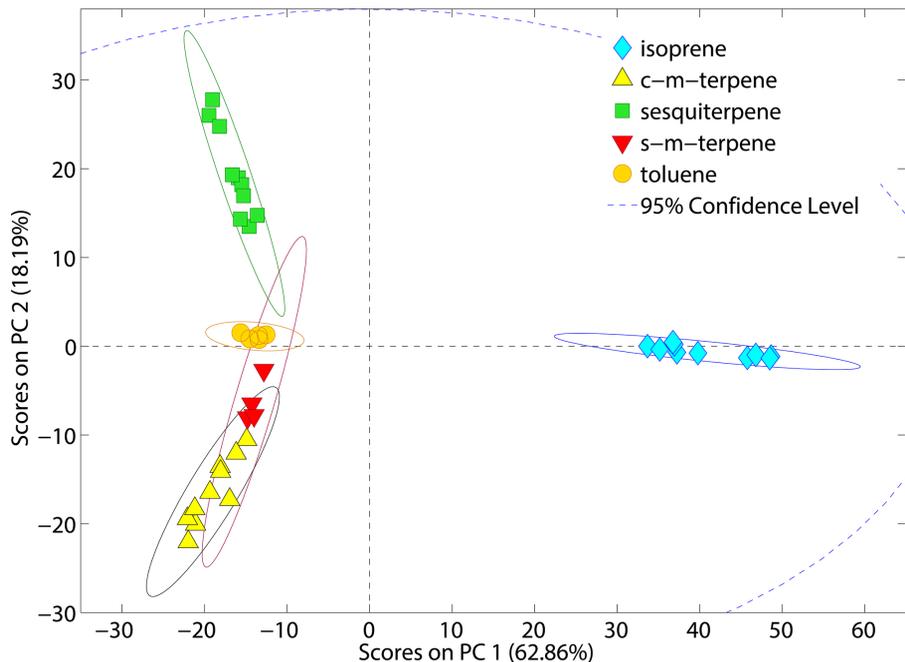
**Figure 3.** PCA scores plot of the second vs. first principal components derived from the PCA analysis of the mesocosm test set using the PCA model developed from the isoprene, cyclic monoterpene ( $\alpha$ -pinene and limonene), sesquiterpene ( $\beta$ -caryophyllene) and straight chain monoterpene (myrcene and linalool) chamber data. Classification confidence limits = 95 %.

**Deconvolution of complex atmospheric datasets**

K. P. Wyche et al.

Title Page	
Abstract	Introduction
Conclusions	References
Tables	Figures
◀	▶
◀	▶
Back	Close
Full Screen / Esc	
Printer-friendly Version	
Interactive Discussion	





**Figure 4.** PCA scores plot of the second vs. first principal components derived from the PCA analysis of the toluene test set using the PCA model developed from the isoprene, cyclic monoterpene ( $\alpha$ -pinene and limonene), sesquiterpene ( $\beta$ -caryophyllene) and straight chain monoterpene (myrcene and linalool) chamber data. Classification confidence limits = 95 %.

Deconvolution of complex atmospheric datasets

K. P. Wyche et al.

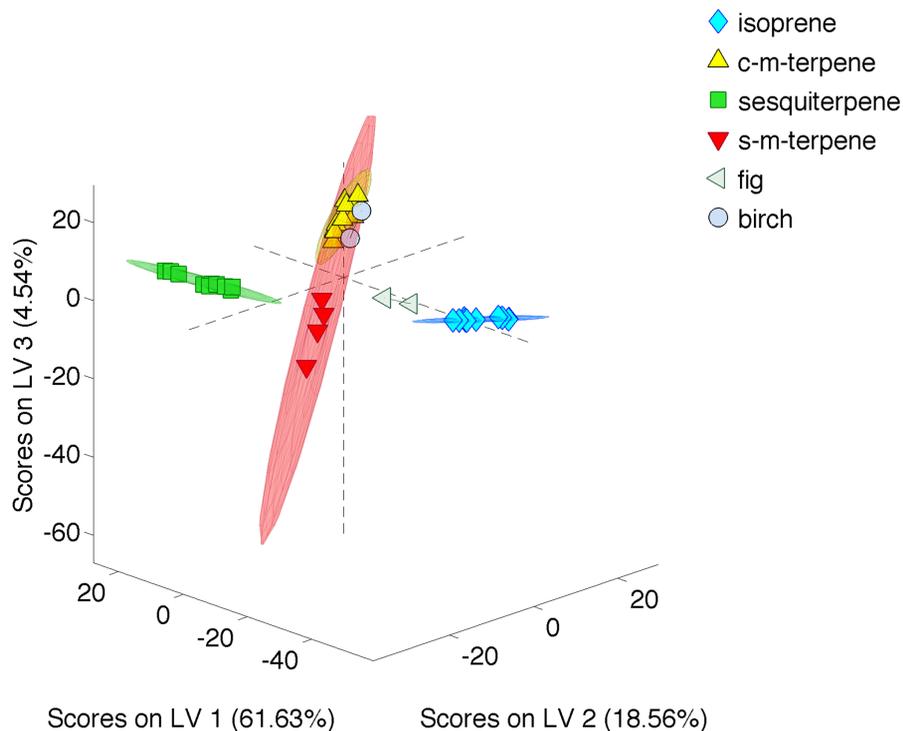
Title Page	
Abstract	Introduction
Conclusions	References
Tables	Figures
◀	▶
◀	▶
Back	Close
Full Screen / Esc	
Printer-friendly Version	
Interactive Discussion	





## Deconvolution of complex atmospheric datasets

K. P. Wyche et al.



**Figure 6.** Scores plot of the first three latent variables derived from the PLS-DA model analysis of the isoprene, cyclic monoterpene ( $\alpha$ -pinene and limonene), sesquiterpene ( $\beta$ -caryophyllene), straight chain monoterpene (myrcene and linalool), fig and birch chamber data. Classification confidence limits = 95 %.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

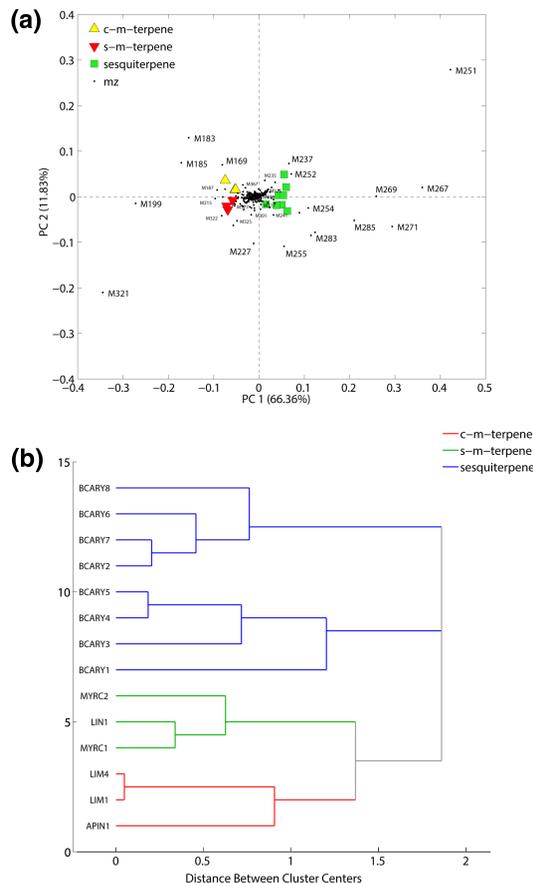
Printer-friendly Version

Interactive Discussion



Deconvolution of complex atmospheric datasets

K. P. Wyche et al.



**Figure 7.** (a) Loadings bi-plot of the second vs. first principal components obtained from the PCA of LC-MS aerosol spectra from a subset of terpene experiments and (b) the corresponding HCA dendrogram.

Title Page

Abstract Introduction

Conclusions References

Tables Figures

◀ ▶

◀ ▶

Back Close

Full Screen / Esc

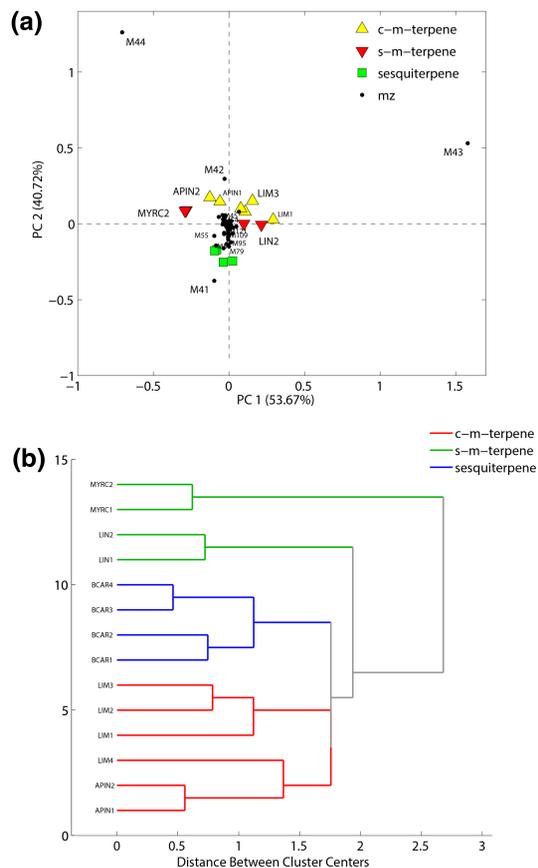
Printer-friendly Version

Interactive Discussion



Deconvolution of complex atmospheric datasets

K. P. Wyche et al.



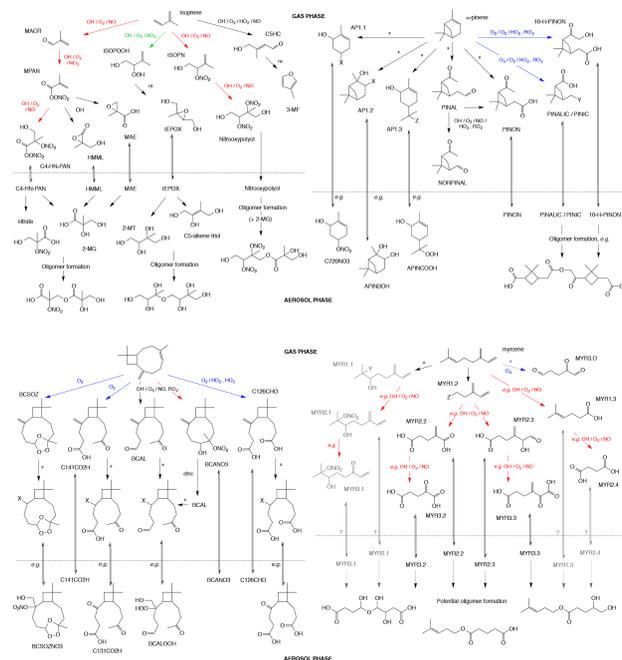
**Figure 8.** (a) Loadings bi-plot of the second vs. first principal components obtained from the PCA of AMS aerosol spectra from of a subset of terpene experiments and (b) the corresponding HCA dendrogram.

Title Page	
Abstract	Introduction
Conclusions	References
Tables	Figures
◀	▶
◀	▶
Back	Close
Full Screen / Esc	
Printer-friendly Version	
Interactive Discussion	



## Deconvolution of complex atmospheric datasets

K. P. Wyche et al.



**Figure 9.** Simplified schematic illustrating some of the important mechanistic pathways in the gas-phase oxidation of isoprene,  $\alpha$ -pinene,  $\beta$ -caryophyllene and myrcene, and the associated mass transfer to the particle-phase. Red arrows and text = “high”  $\text{NO}_x$  pathways, green arrows and text = “low  $\text{NO}_x$ ” pathways, blue arrows and text = ozonolysis reactions, grey arrow and text = speculative, dashed arrows = multiple steps. \* = multiple photooxidative routes initiated by reaction with OH (i.e. involving the reactants – OH,  $\text{O}_2$ , NO,  $\text{HO}_2$  and/or  $\text{RO}_2$ ), leading to structurally similar products containing different functional groups.  *$\alpha$ -pinene mechanism* – X = OH, =O, OOH or  $\text{ONO}_2$ ; Y = CHO or  $\text{C}(\text{O})\text{OH}$ ; Z = OH, OOH or  $\text{ONO}_2$ .  *$\beta$ -caryophyllene mechanism* – X =  $\text{CH}_2\text{OH}(\text{OH})$ ,  $\text{CH}_2\text{OH}(\text{OOH})$ ,  $\text{CH}_2\text{OH}(\text{ONO}_2)$  or =O. *Myrcene mechanism* – Y = OOH or  $\text{ONO}_2$ ; Z = CHO or  $\text{C}(\text{O})\text{OH}$ . See text, Sect. 5 for references.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

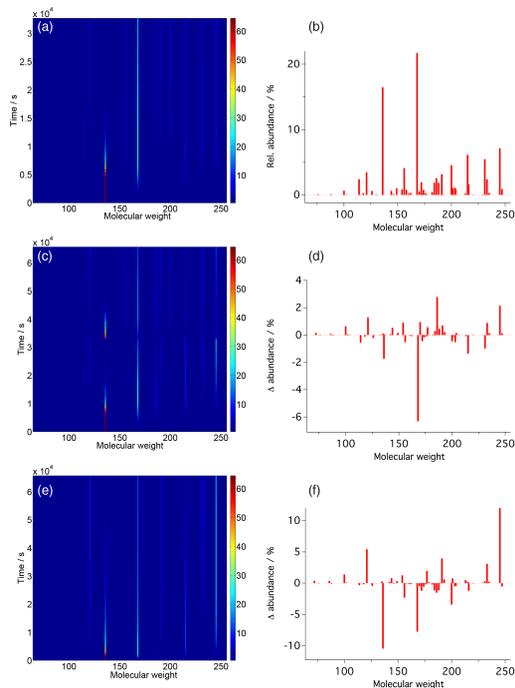
Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion





**Figure 10.** Results from MCM  $\alpha$ -pinene photooxidation simulations. **(a and b)** basic  $\alpha$ -pinene photooxidation; **(c and d)** spiked injection of  $\alpha$ -pinene, continuous HONO input; **(e and f)** continuous  $\alpha$ -pinene and HONO input. Left hand image plots show the evolution of the respective systems over the molecular weight region of interest with time; colour scale = relative abundance/%. Right hand plots **(b)** relative abundance of simulated molecular weights during straight  $\alpha$ -pinene photooxidation; **(d)** difference in relative abundance of simulated molecular weights between double injection of  $\alpha$ -pinene continuous HONO input and straight  $\alpha$ -pinene photooxidation; **(f)** difference in relative abundance of simulated molecular weights between continuous  $\alpha$ -pinene and HONO input and straight  $\alpha$ -pinene photooxidation. See text for details.